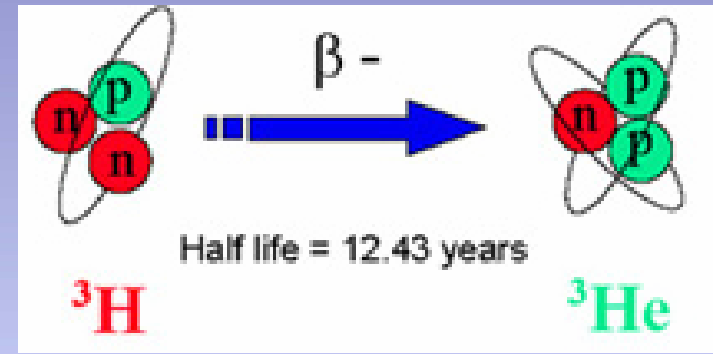


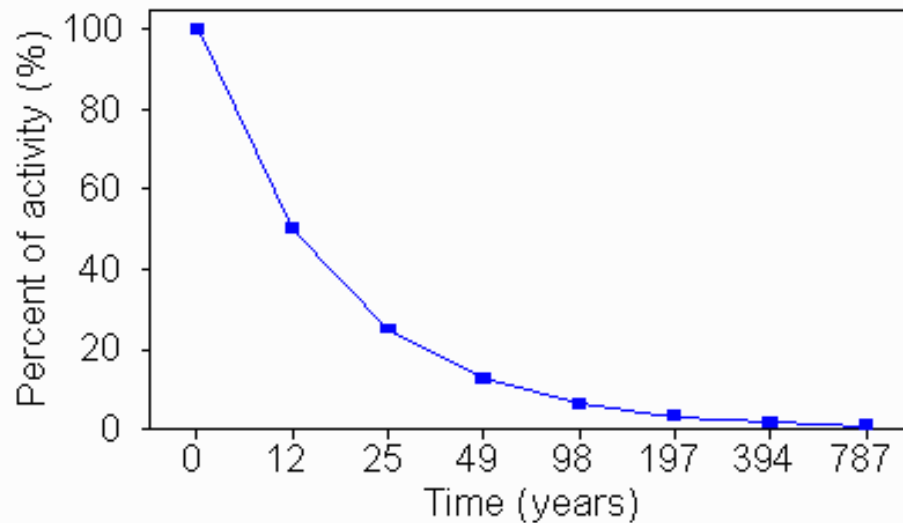
Groundwater age dating

${}^3\text{H} \rightarrow {}^3\text{He}$ A good tracer because ${}^3\text{H}$ is right in the water. HTO instead of H_2O .



Decay Curve for Tritium

Amount of radioactivity vs time



Tritium forms ${}^3\text{He}$ by beta decay with a half-life of about 12.4 years.

$$1\text{TU} = 1\text{T} / 10^{18}\text{H}$$

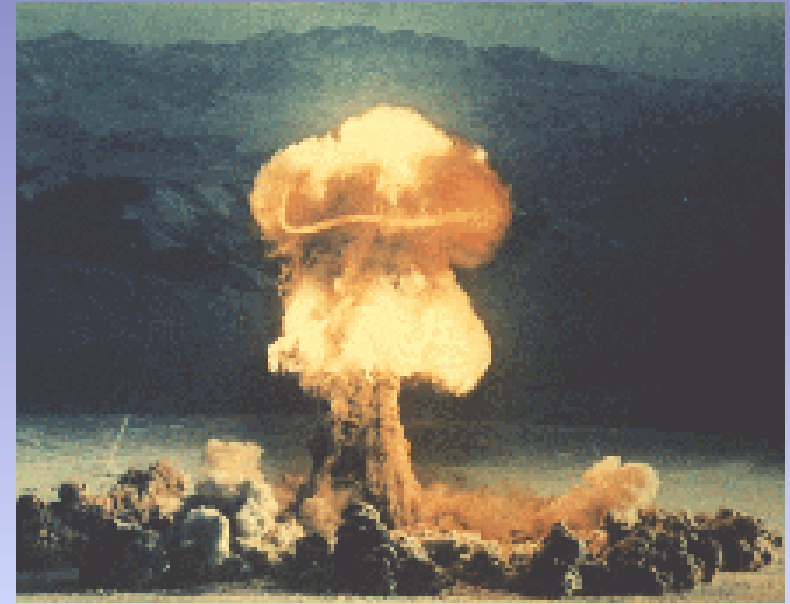
$${}^3\text{He}/{}^3\text{H} = \text{age.}$$

- Tritium has been used for many years (produced by atmospheric nuclear testing). Nobody tried to apply the decay pair in ground water until the late 1980's (although it had been suggested by the oceanographers that were already using it 23 years earlier- more concerned about water supply, agriculture).

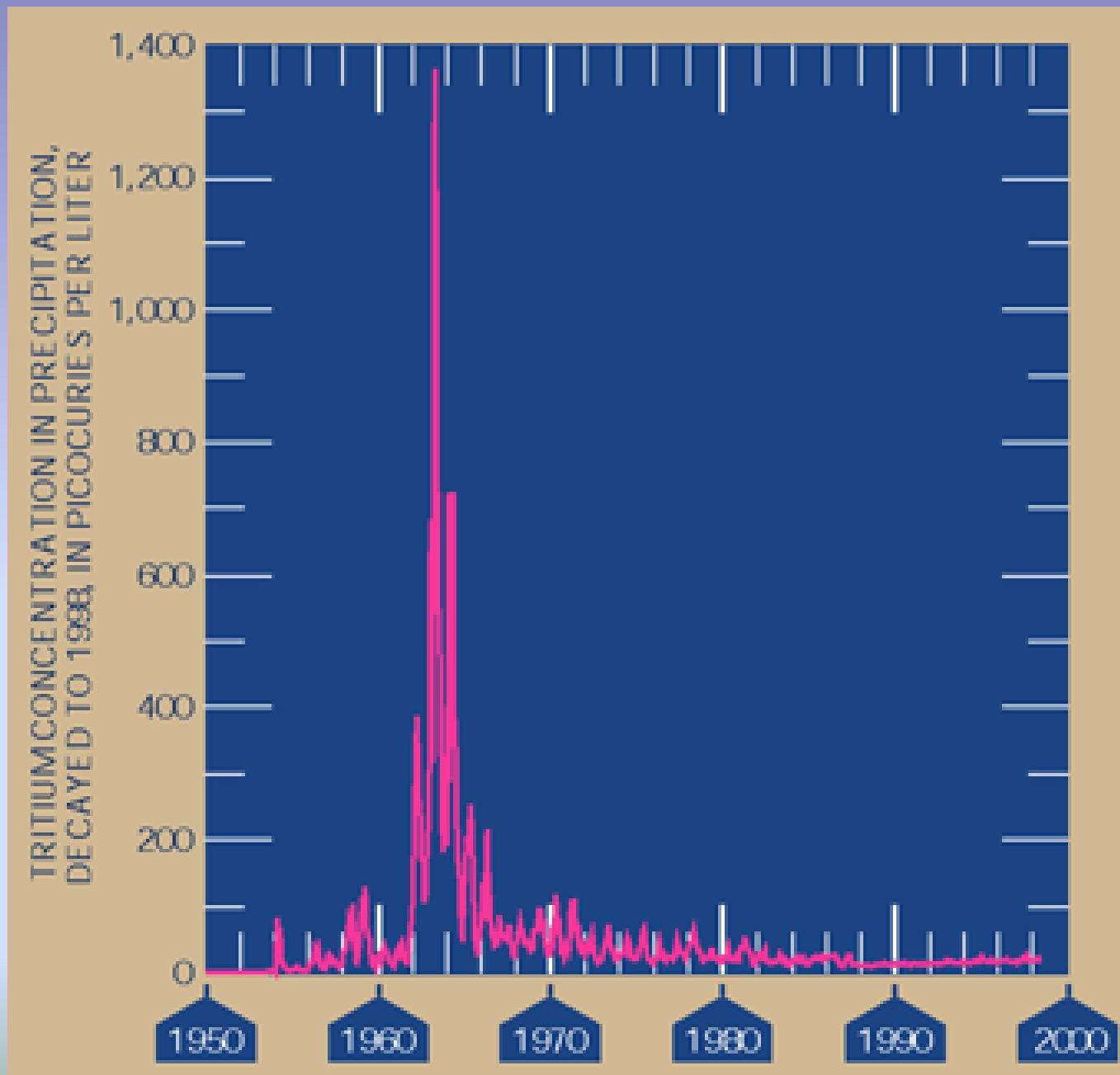
* This method is not seen in any textbook and therefore hasn't caught on that fast. Many hydrologists approach from physical rather than a chemical viewpoint, but the chemical can tell a lot about the system.

* **Dramatic increase in tritium as the Cold War heated up.** At its peak, the furious explosions of nuclear weapons prior to the ban-- mostly Russian tests (others may have gone out over the oceans).

- About 25% of tritium in troposphere mixes into the stratosphere each year-- reason for gradual decrease with time.
- **Last 20 years or so-- stabilized at about 30 tritium units.** The tritium is still in the hydrologic system → due to weapons, plants, etc?
- The dismantling of nuclear war heads may cause an increase in tritium again. And a lot of tritium is contained in the H-bomb which must be refurbished in tritium about every 18 months due to leakage.



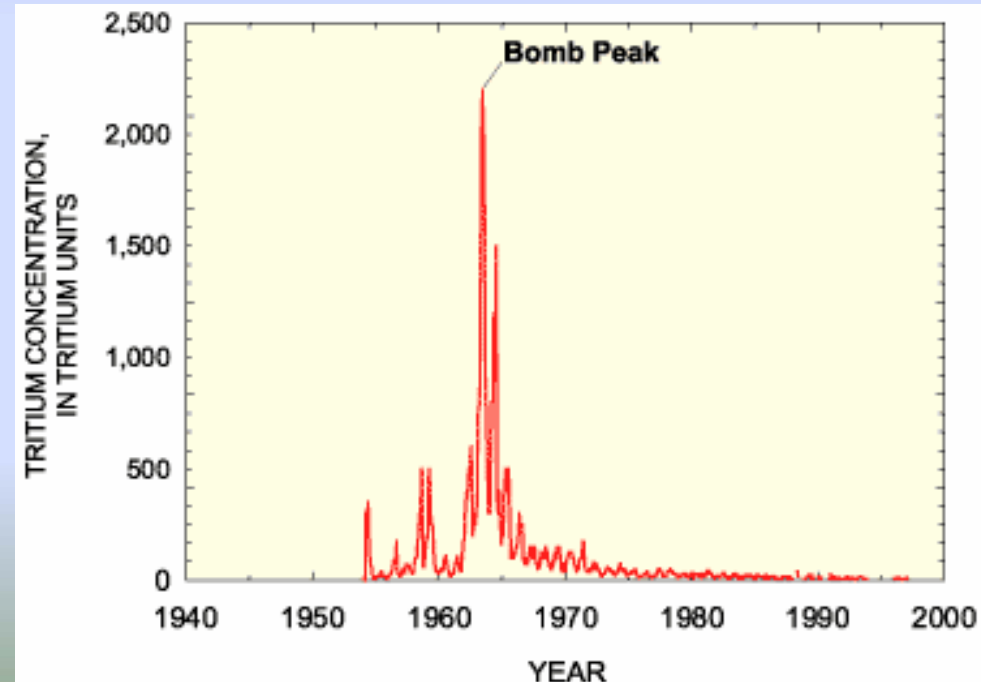
Tritium Concentration in Precipitation, Decayed to 1998



- **Places are affected differently by the tritium** → Places with large marine influence have lower tritium values while places that are land-locked with mainly only local precipitation, have high ratios. Mexico has only 3-4 tritium units.

Latitude: 50°N → high
0° → lower
30°S → lower

Estimated tritium concentration in precipitation for the Tidewater area of Virginia, 1954 to 1997.



Data derived from estimation technique of Michel (1989). Values not corrected for radioactive decay.

http://images.google.com/imgres?imgurl=http://toxics.usgs.gov/photo_gallery/photos/tritium/cab_fig18a.gif&imgrefurl=http://toxics.usgs.gov/definitions/tritium.html&h=325&w=340&sz=21&tbnid=xJJSz6yBEsgJ:&tbnh=109&tbnw=114&start=1&prev=/images%3Fq%3Dtritium%2Batmosphere%26hl%3Den%26lr%3D

→ function of where the input was and where the stratosphere mixes (more readily at high latitudes).

Idea is to take a known system and to try new idea to see if the results are the same (new idea → tritium-helium system).

Tritium/Helium technique

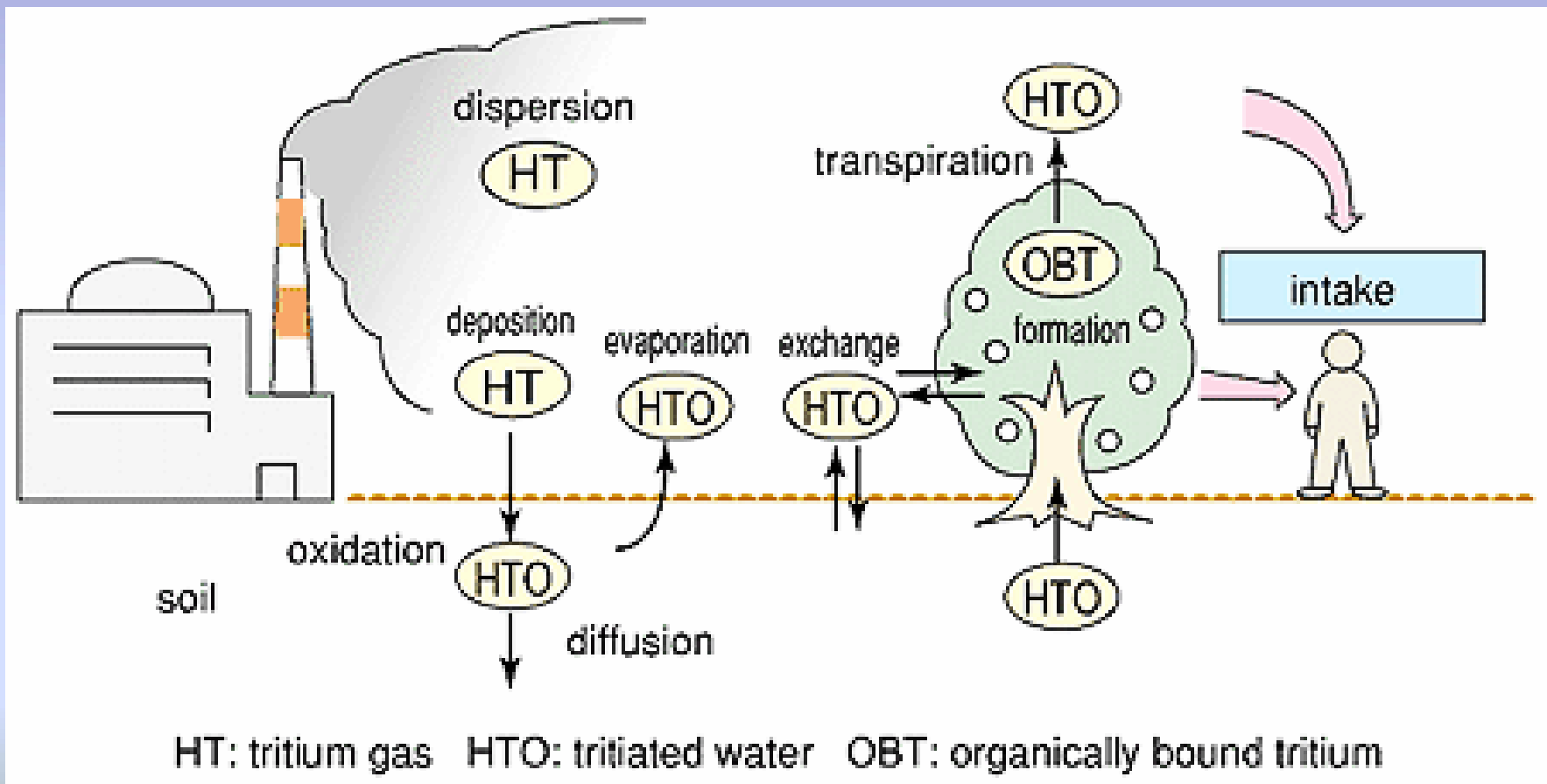
^3He → normally 7 ppt (parts per trillion) in the atmosphere.

^4He → normally 5.2 ppm (parts per million) in the atmosphere
(i.e. much more common).

When rainwater falls to the earth ---> it is in solubility equilibrium with the atmosphere and therefore some of the ^3He and the ^4He will dissolve into the ground water at about $68 \times 10^{-12} \text{ cc}^3\text{He}/\text{kg}$ water or about 1700 million atoms He/kg water.

Therefore, have to subtract atmospheric ^3He because it is not produced by decay.

Environmental behavior of tritium after tritium gas was released to the atmosphere



$$t(\text{travel time}) = (1/\lambda) \ln\left(\frac{{}^3\text{He}^*}{{}^3\text{H}} + 1\right)$$

$\text{He}^* = \text{He}$ from tritium decay

$1/\lambda = \text{decay constant}$

$$N = N_0 e^{-\lambda t} \quad \rightarrow \quad {}^3\text{H} = {}^3\text{H}_0 e^{-\lambda \tau} \quad \rightarrow \quad {}^3\text{H} = ({}^3\text{H} + {}^3\text{He}) e^{-\lambda \tau}$$

$$[{}^3\text{He}]_w = kP_{{}^3\text{He}} \quad \rightarrow \quad ({}^3\text{He} / {}^4\text{He})_w = 0.985 ({}^3\text{He} / {}^4\text{He})_{\text{air}}$$

Things to note:

- This travel time doesn't require you to know the initial tritium value
- ${}^3\text{He} + {}^3\text{H} = \text{conservative tracer in ground water.}$
- ${}^3\text{He}/{}^3\text{H}$ 'clock' is set to zero at water table.
- Why? At water table in contact with the atmosphere- in solubility equilibrium where ${}^3\text{He}/{}^3\text{H} = 0$.

Time here also has a nice progression from surface to deep layers and parallels the travel time based on Cl. **Why is it offset?**
Because the clock was set to zero at about 70-80 cm below the water table (about 1 year's travel time). This is because the water travels in pulses and usually mostly in the springtime.

Age gradient ---> a very robust establishment of vertical velocity.

2 m/5 years = $V_z = 40$ cm/year.

$(V_z)(\text{Porosity}) = 40(.3) = 12$ cm/year = recharge.

Vertical age gradient tells you how much recharge is coming through into the system → how fast things are moving

What determines the flow, is what you put in at the top.

- With rainfall = 80-100 cm/year, the recharge at the source is about 60 cm/year.
 - but 1 km away, recharge is only 12 cm/year. Why? Because the source, as a dump is relatively devoid of any life, whereas the trees 1 km away extract 80-90% of the water.
 - How do you get horizontal and vertical velocities?
 - Might want to know how quickly a contaminant will migrate off sight
 - Well is screened over entire interval to get the most information in the radial sense.
- A lot of information is lost about the aquifer by sampling over entire interval.**
- Don't know exactly where the water is coming from.
 - should only sample about 10 centimeters instead of 1-2 meters.

If trying to find contaminant and it is moving at a narrow level, it will enter the well and be diluted by relatively pure water. "Water table aquifer"

^3H - ^3He dating of ground water

Cape Cod, MA

Purpose:

1. Demonstrated the usefulness of this technique in establishing rates of ground water flow.
2. Assists in establishing rates of plume migration.

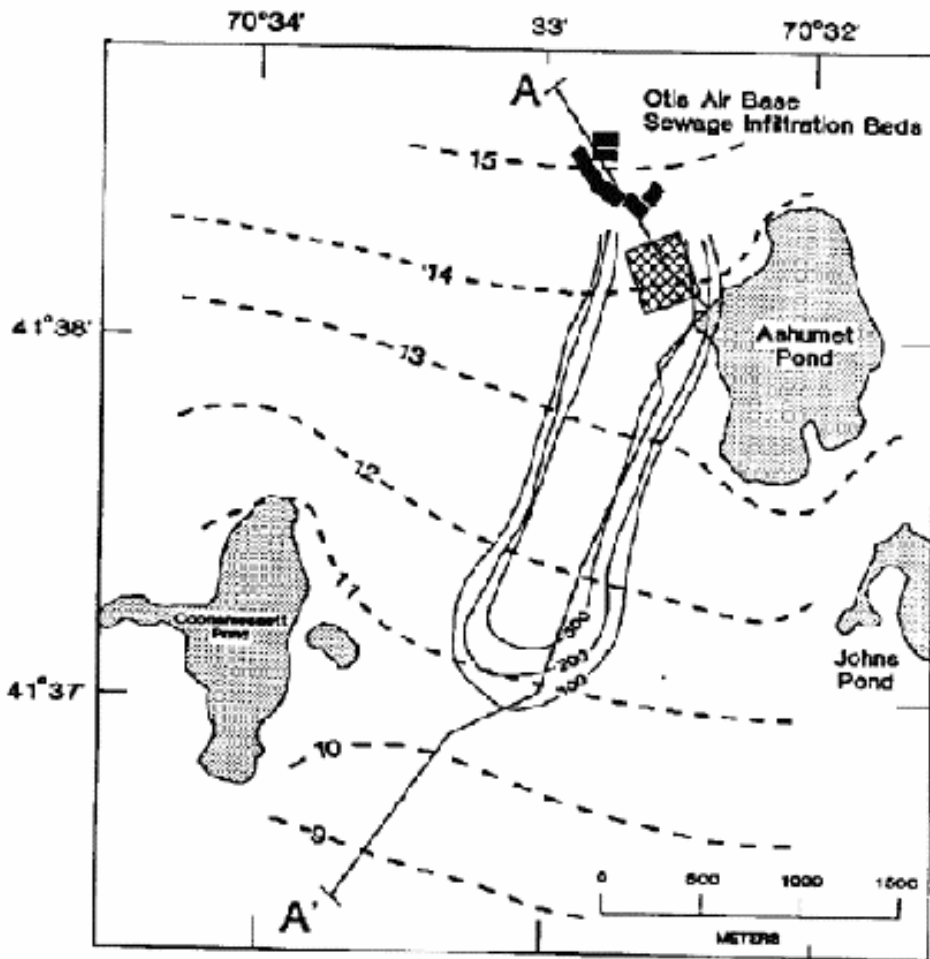
- Tritium (^3H) input from thermonuclear weapons tests (peak in 1963); no longer a sustained input.

- Age doesn't require knowledge of tritium input function.





- Amount of tritium in 1 nuclear warhead = amount in the earth's atmosphere.



Munitions uncovered in a burial site at the J-1 Range during IAGWSP investigations.



EXPLANATION

-  **TRACER-TEST SITE** — Shows area in figure 4.
-  **LINE OF EQUAL BORON CONCENTRATION, 1978-79**—
Concentrations in micrograms per liter.
-  **WATER-TABLE CONTOUR, NOVEMBER 1979**—
Altitude of water table in meters. Datum is sea level.
-  **LINE OF SECTION SHOWN IN FIGURES 2 AND 14.**

LeBlanc et al. (1991) WRR, 27, 895

${}^3\text{H}$ ----(beta decay β^-) \rightarrow ${}^3\text{He}$ (trace isotope of He)
Half life : $t_{1/2} = 12.4$ years

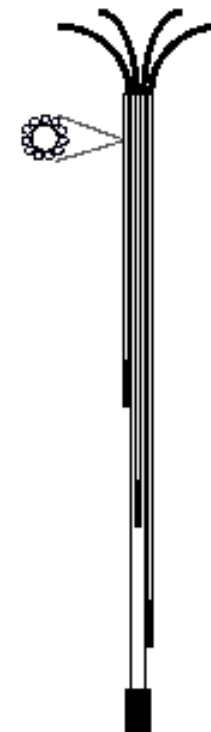
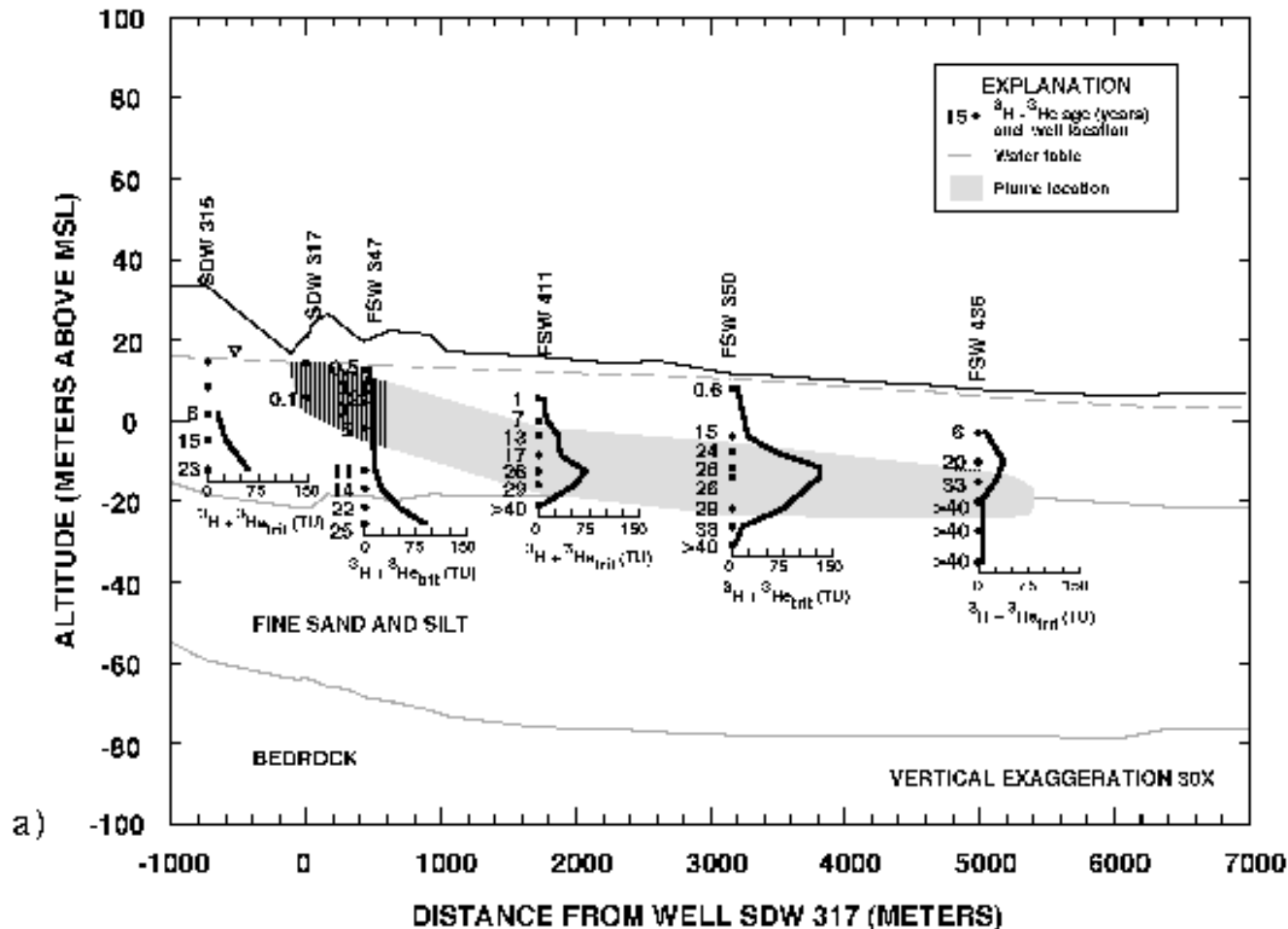
Assume all ${}^4\text{He}$ in atmosphere, correct for ${}^3\text{He}$. ${}^3\text{He}$ less soluble in water than ${}^4\text{He}$, so any excess ${}^3\text{He}$ is from tritium (${}^3\text{He}^*$) [Solubilities: ${}^3\text{He}$ - about 7 ppt ; ${}^4\text{He}$ - about 5.2 ppm]

$$\text{-- i.e. } {}^3\text{He}_{\text{total}} = {}^3\text{He}_{\text{sol}} + {}^3\text{He}^*_{\text{trit}}$$

Travel time $t = (1/\lambda) \ln [({}^3\text{He}/{}^3\text{H}) + 1]$

We want to measure the excess over solubility. Assume ${}^4\text{He}$ in atmosphere, assume ${}^3\text{He}/{}^4\text{He}$ in air is constant = 1.4×10^{-6} .

Cape Cod - $^3\text{H}/^3\text{He}$ dating



The $^3\text{He}/^3\text{H}$ clock is set to zero at the water table because vadose zone exchange with atmosphere through change in temperature (air exchange) -- solubility equilibrium.

Vertical Age Gradient = recharge rate (combine recharge rate and hydraulic head)

FS-12 Field Site with Water Level Contours, the location of multilevel wells, and the extent of a benzene plume determined in the 1991 study.

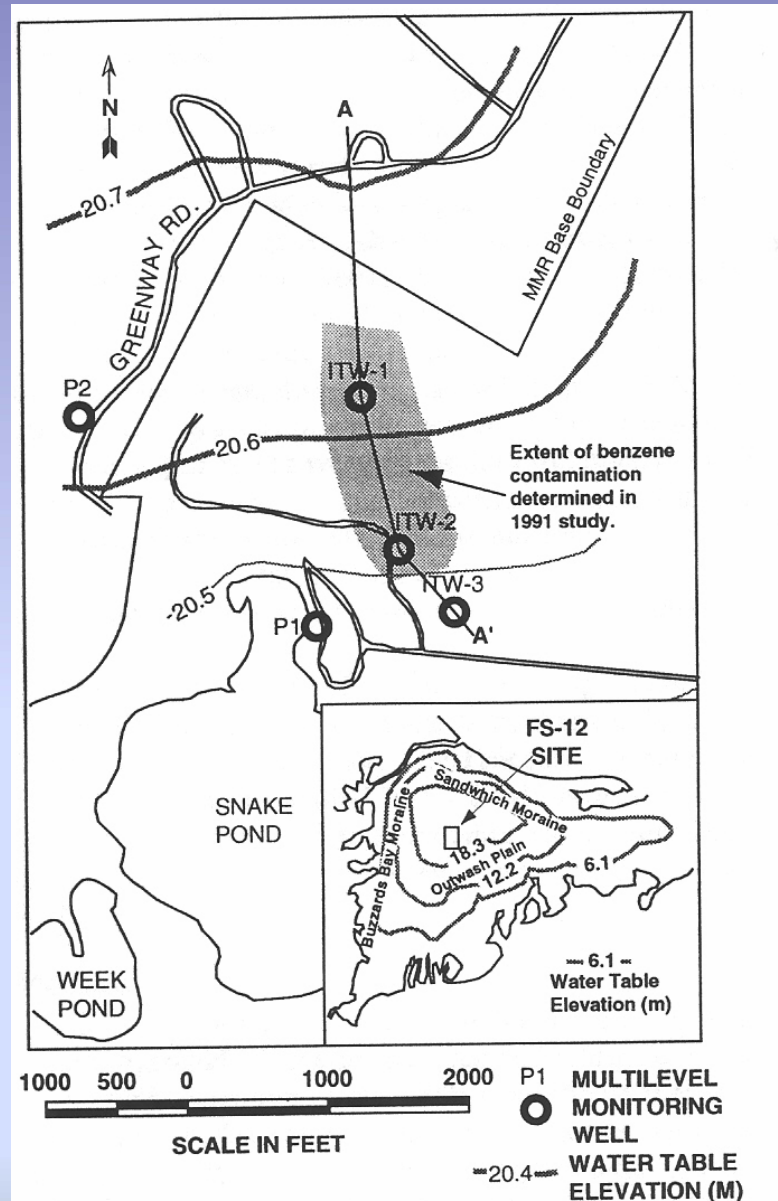


Fig. 1. FS-12 field site with water level contours, the location of multilevel monitoring wells, and the extent of a benzene plume determined in the 1991 study.

- ^3He and ^3H is a conservative ground water tracer because there is no chemistry involved (He is inert); movement is controlled only by physical mechanisms.
- Helium diffusion is faster than tritium diffusion so is complicated by diffusion effects at peak age.
- When you sample below the water table, have to screen over a very small interval (<15cm).
 - Ages are vertically stratified.
 - But, most regulations make you screen over 5 feet!
 - EPA says well should be > 2 feet.
- Bad because there's a large interval, then remove a lot of water from the system to clean the well.
- Must remove 3 well volumes until you can be sure that your water is really up into the aquifer.
- Also- screening over a large interval means getting several ages at once.
- To sample only over a discrete interval, place small wells in the hole at different depths.

Cape Cod is a terminal moraine with outwash plain (mostly sand).

Snake Pond is a surface expression of the water table, as are most lakes.

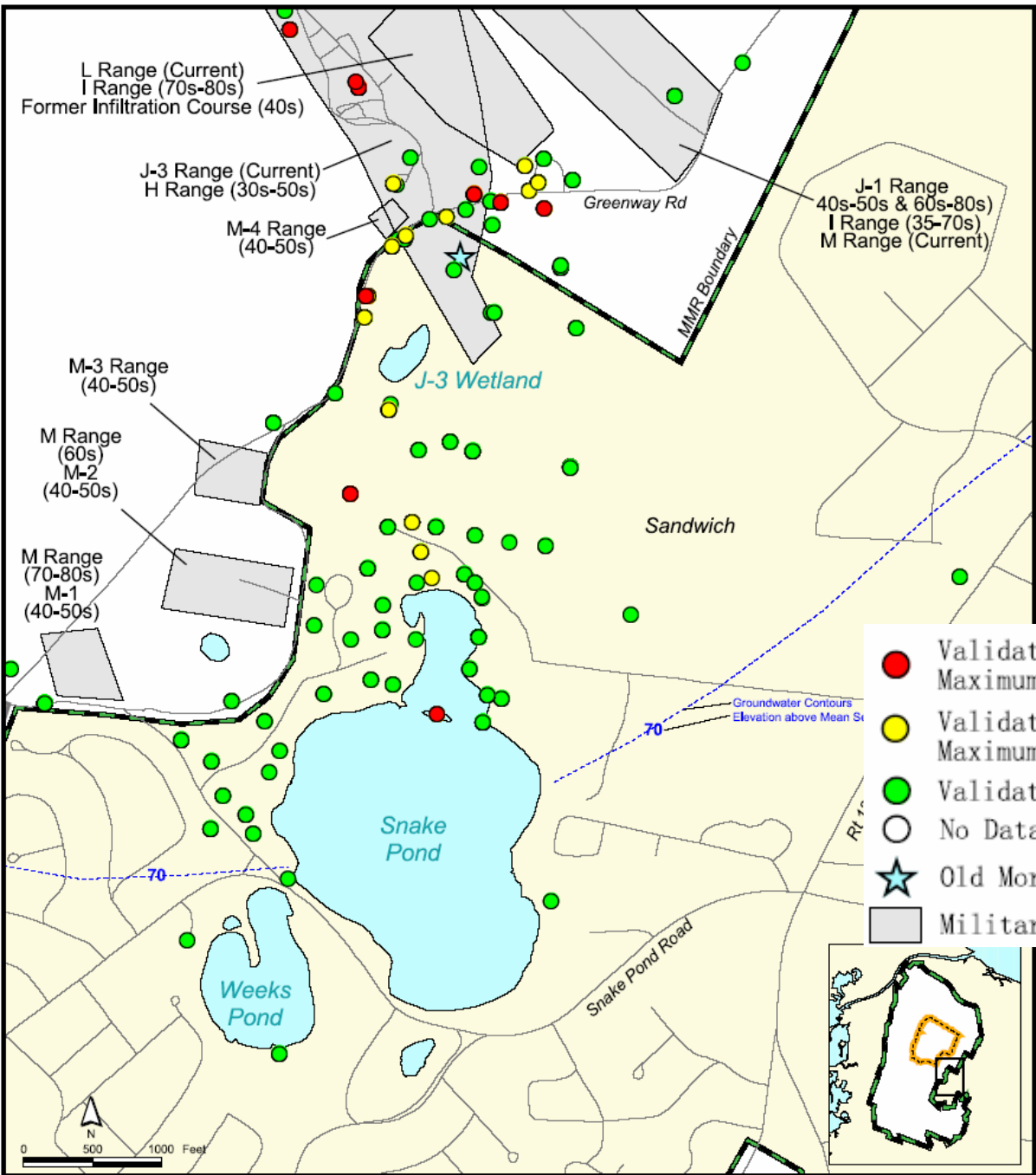
Snake Pond

Spill components at this base

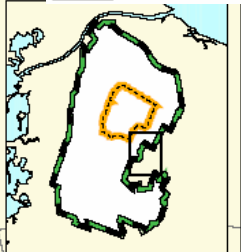
1. Aviation fuel.
2. JP4 (jet fuel; more like kerosene).



IAGWST Explosives and Perchlorate in Groundwater Compared to Maximum Contaminant Level/Advisories



- Validated Detection Greater Than or Equal to Maximum Contaminant Level/Health Advisories
- Validated Detection Less Than Maximum Contaminant Level/Health Advisories
- Validated Non-detect
- No Data Available
- ★ Old Mortar Position
- Military Ranges



Validated Data as of 02/14/02

Looked at xylene, benzene, EDB (= ethylene dibromide, volatile, in Pb gasoline, keeps Pb from Pb oxide).

- xylene, benzene - only sparingly soluble in ground water and are large molecules (→ move slowly)
 - EDB - small molecules, moves at about the rate of ground water.
- ***vertical stratification in ground water.
 - The major spill of aviation gas in 1974, which sank quickly to the water table even though less dense than water.
 - * a peak of all 3 in 1976.

Profiles of $^3\text{H}/^3\text{He}$ and $^3\text{H} + ^3\text{He}$

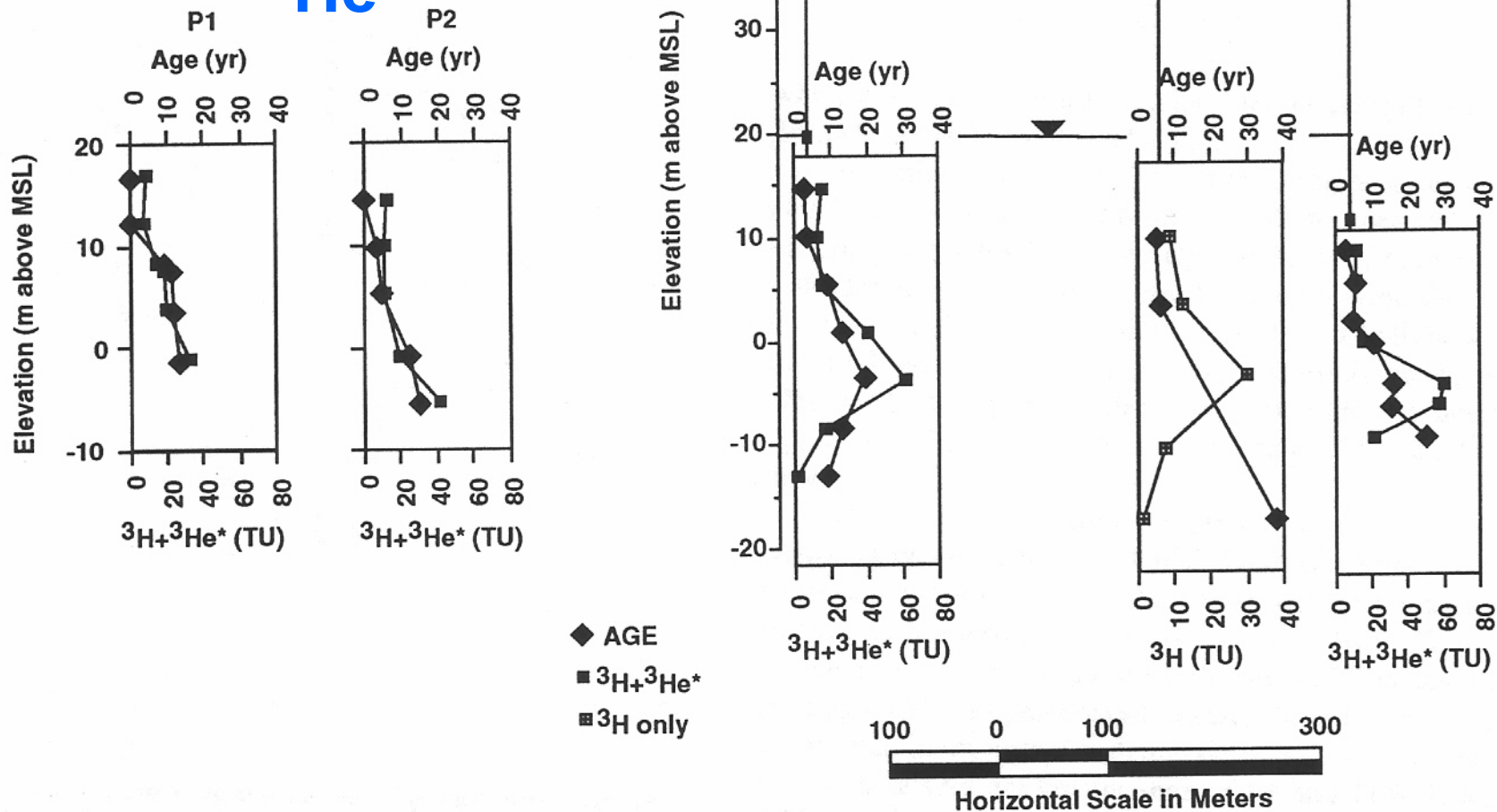


Fig. 2. Profiles of $^3\text{H}/^3\text{He}$ age and $^3\text{H} + ^3\text{He}^*$. The cross section from ITW-1 to ITW-3 is oriented along a ground-water flow line. Note that for ITW-2, only ^3H is plotted because a complete profile of $^3\text{He}^*$ was not obtained. Where duplicate samples were obtained, the plotted values are averages with outliers removed.

Profiles of organic contaminants and ground-water age

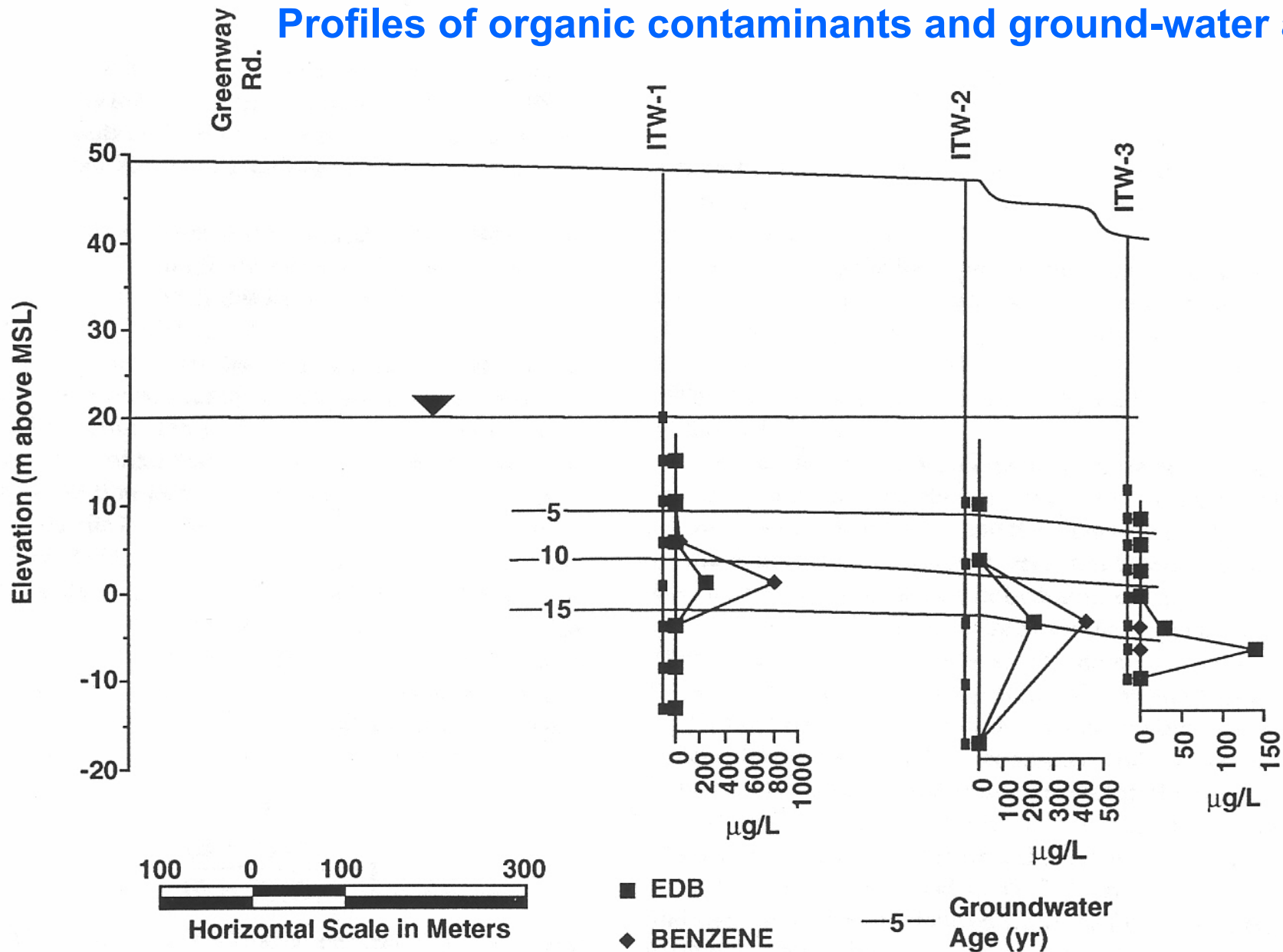


Fig. 3. Profiles of organic contaminants and ground-water age. The cross section from ITW-1 to ITW-3 is oriented along a ground-water flow line. **Solomon et al., 1995**

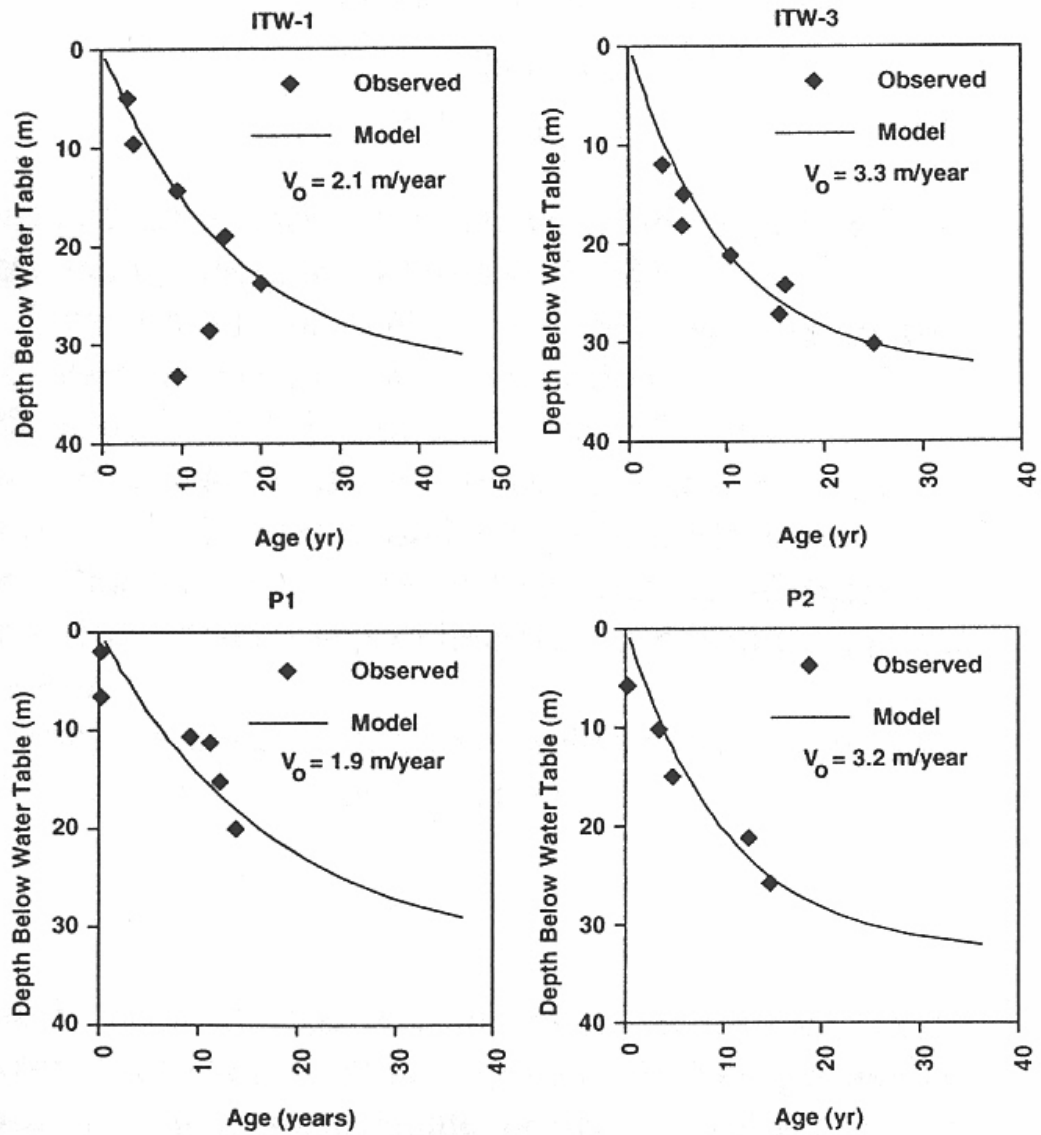


Fig. 4. Profiles of observed and simulated ground-water age. The simulated profiles were obtained by optimizing the fit between the observed and simulated age by adjusting the vertical velocity at the water table.

Profiles of Observed and Simulated Ground-Water Age

Simulated stream function, hydraulic potentials, and the configuration of the water table

Solomon et al., 1995

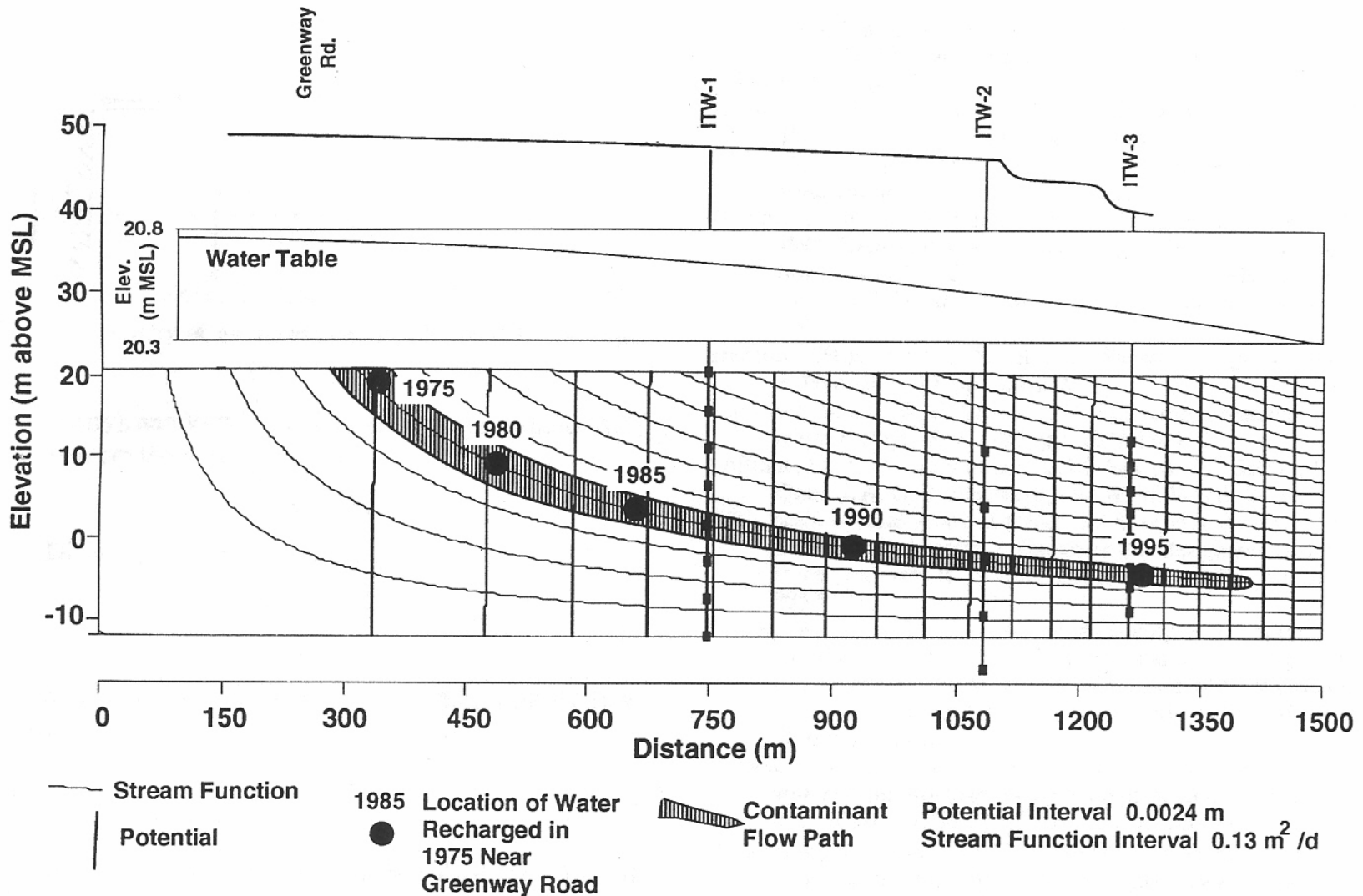
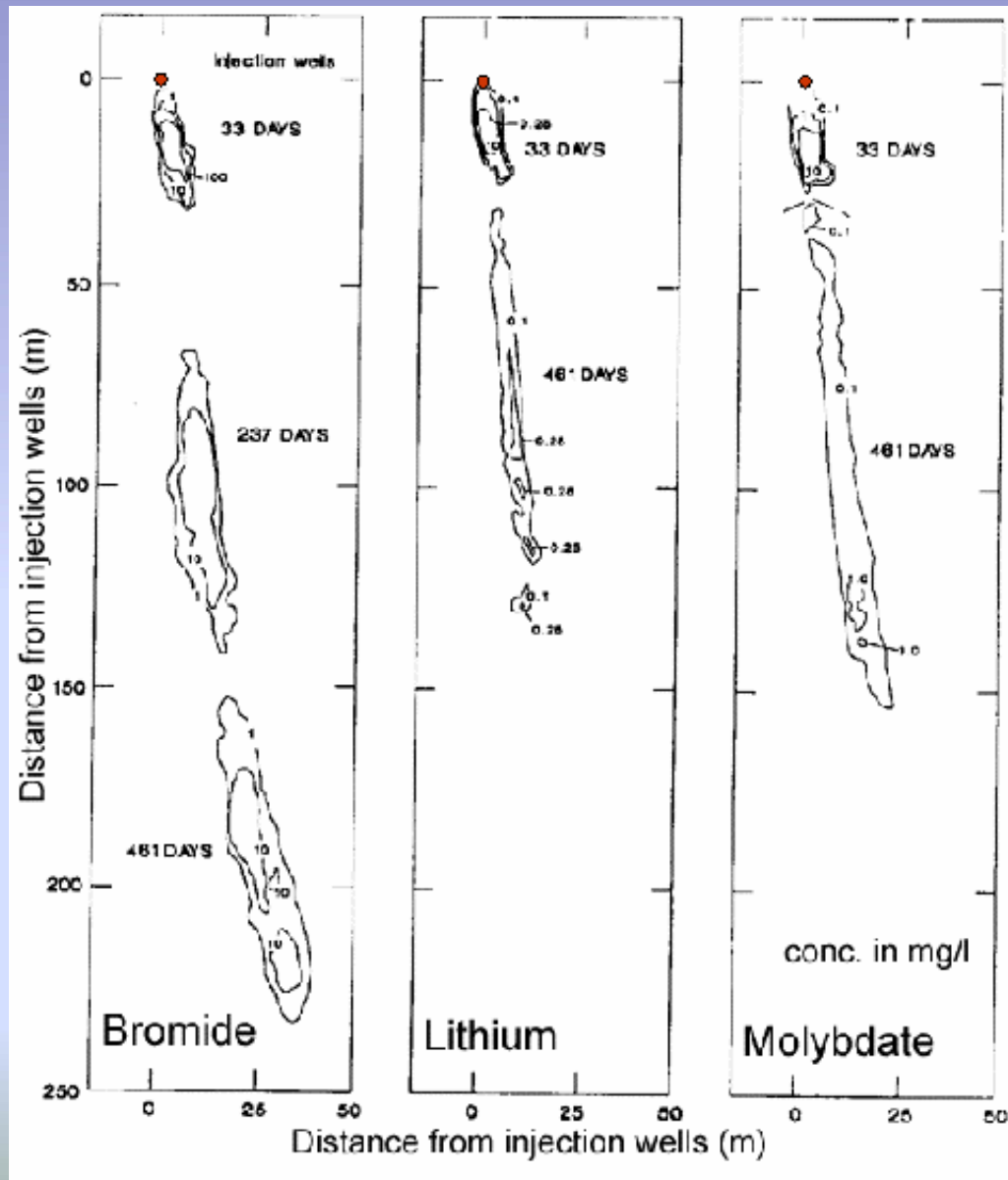
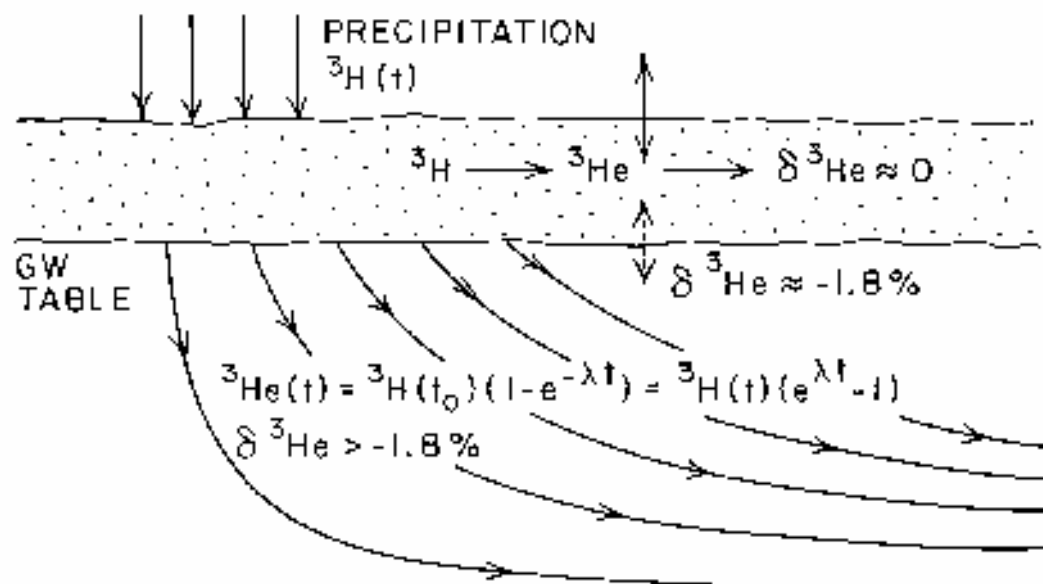
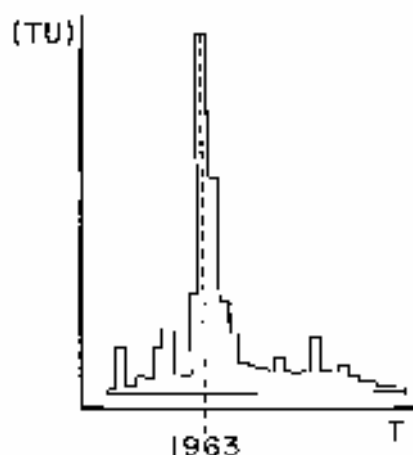


Fig. 5. Simulated stream functions, hydraulic potentials, and the configuration of the water table (note expanded scale). Also shown is the simulated location of water recharged in 1975 beneath Greenway Road at various points in time along the flow system. The shaded region highlights a flow line and does not represent a simulated contaminant plume.

Cape Cod (Le Blanc et al. (1991))



$^3\text{H}/^3\text{He}$ dating



$$\tau = \frac{T_{1/2}}{\ln 2} \cdot \ln \left[1 + \frac{[^3\text{He}]}{[^3\text{H}]} \right]$$

Wetting Fluid - fluid that coats grain surfaces due to surface tension.

- aviation gas not soluble in water and therefore is excluded from water.
 - moves through pore spaces to water table where it floats, sits and slowly dissolves.
- EDB is conservative, is right there with bomb peak when xylene and benzene at 0 (they're already gone).
- * As go deeper, flow speeds up in horizontal away from flow divide.
 - * close to flow divide, longer time spent in vertical as a result of the hydraulic potentials.

Hydraulic conductivity K

- Can look at it in terms of $V_{\text{horizontal}}$ or K_h .
- the vertical component of the flow really shows how the system is behaving.

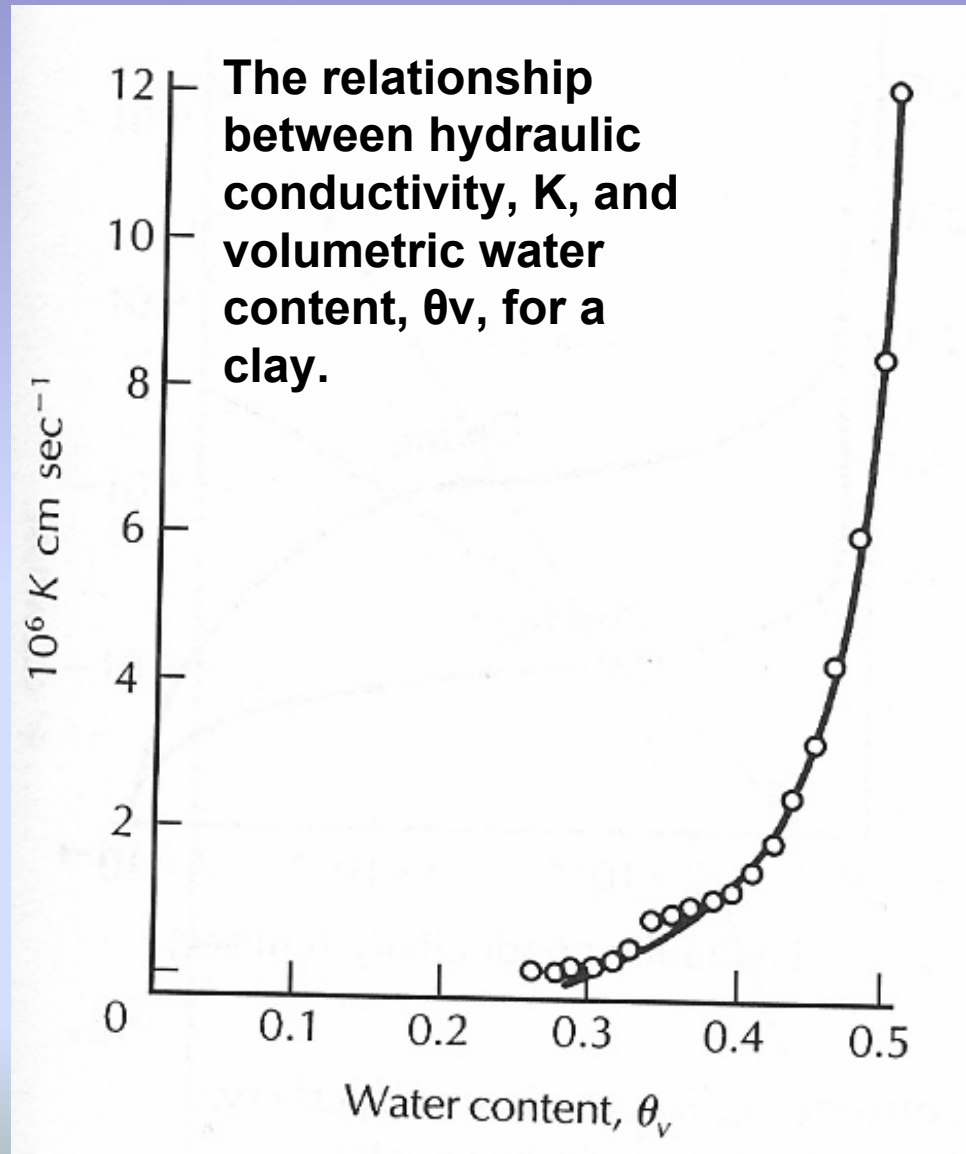
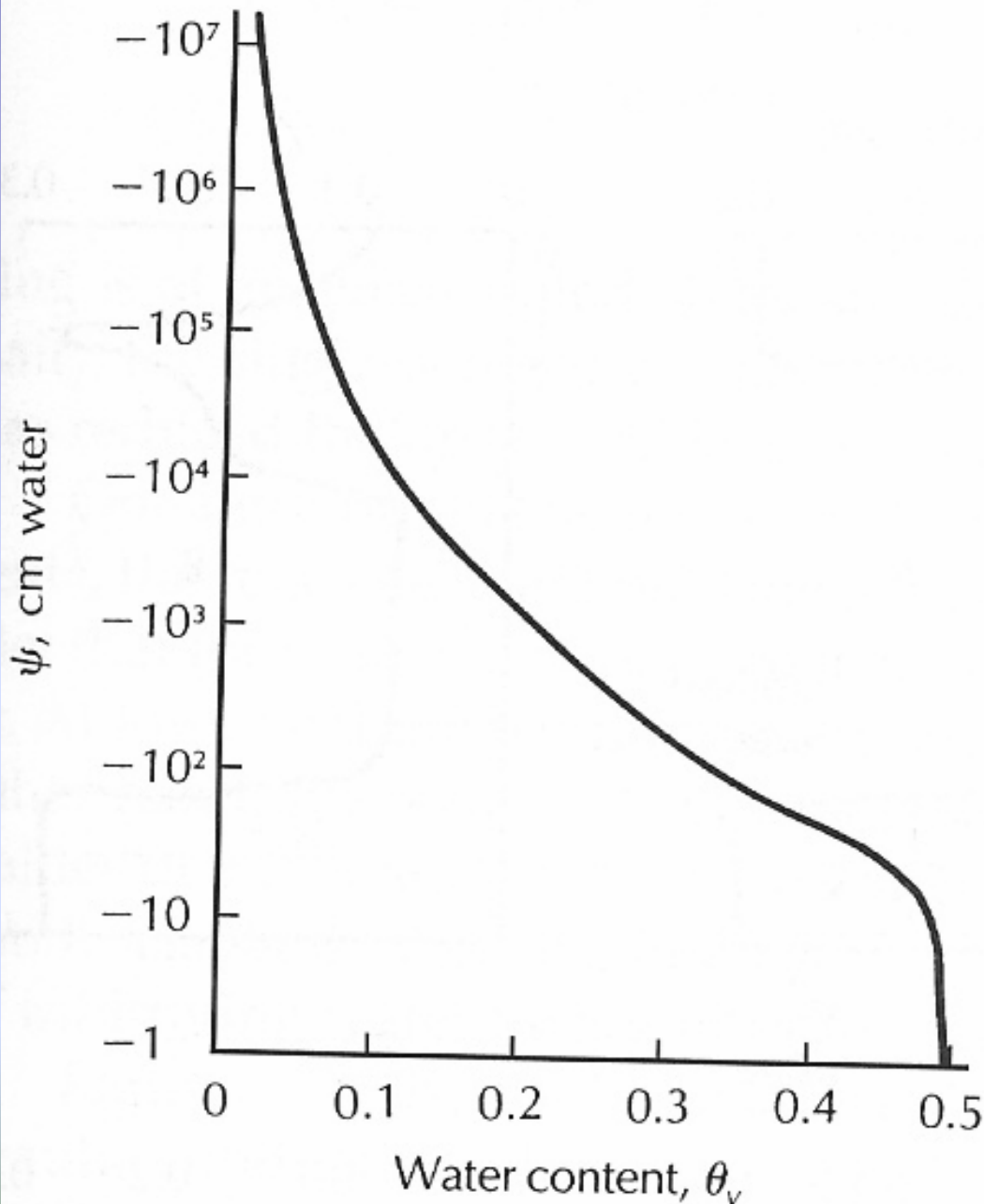
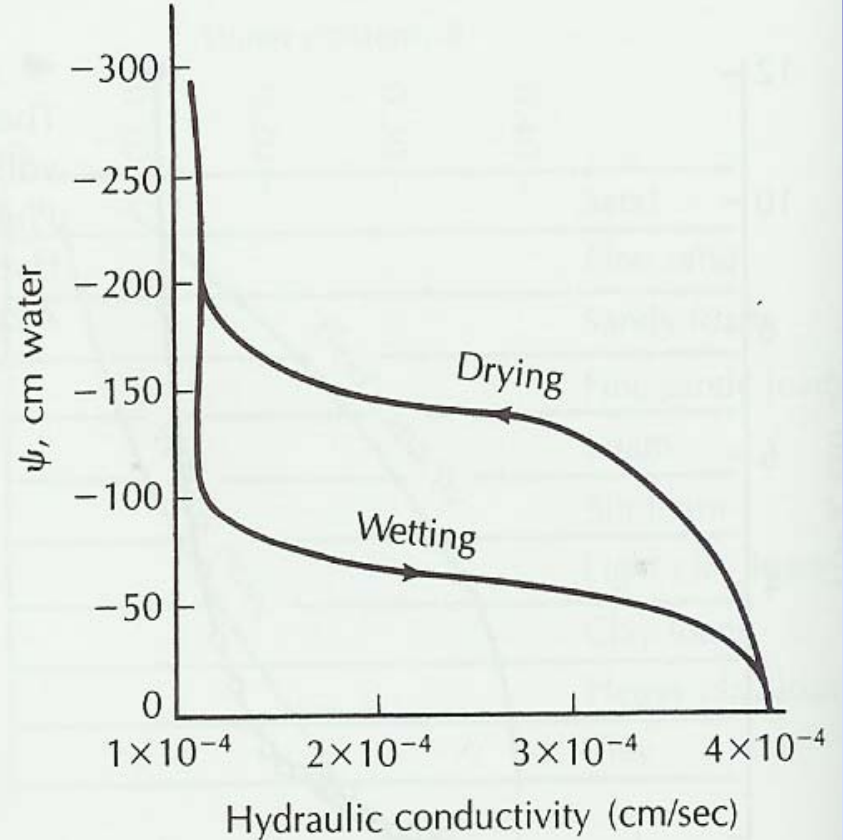
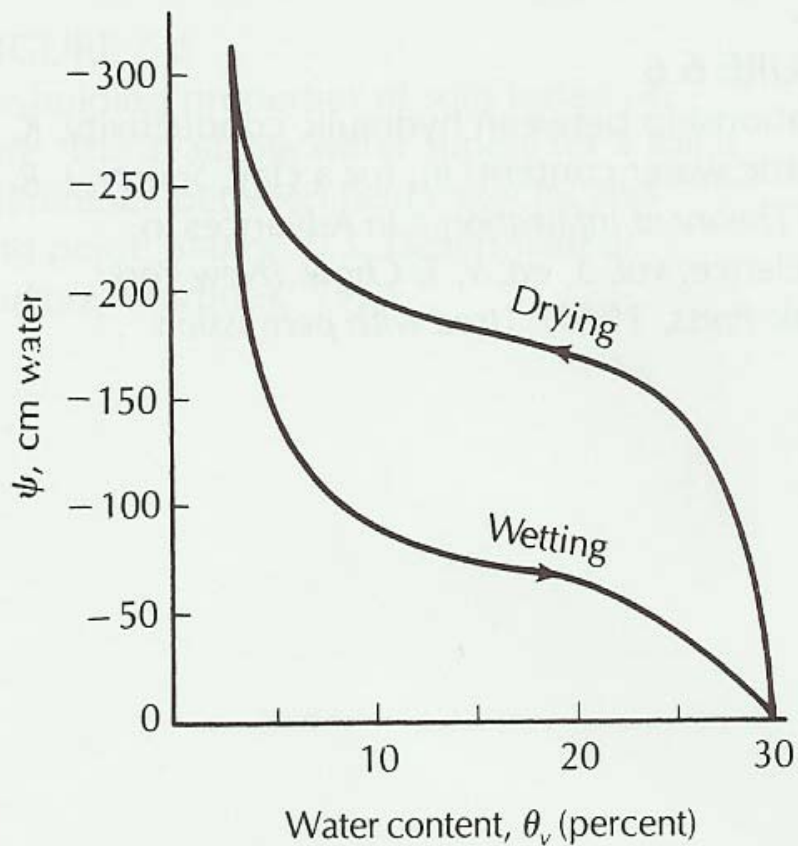


Figure 6.6 Fetter, *Applied Hydrogeology 4th Edition*



The relationship between moisture potential, Ψ , and volumetric water content, θ_v , for clay soil.

Figure 6.7 Fetter, *Applied Hydrogeology 4th Edition*



▲ FIGURE 6.8

Idealized curves showing relationships of volumetric water content, θ_v , hydraulic conductivity, K , and soil-moisture tension head, ψ . The effect is included for wetting and drying cycles.

Hydraulic conductivity K

Vogel Model:

$$K_V = 0.1 K_H$$

K=hydraulic conductivity

$$V_z = \frac{B}{t} * \ln(B/(B-Z))$$

$$V_x = (V_z * x) / B$$

B=aquifer thickness

x=distance from
flow divide

t=travel time

Z=depth below
water table

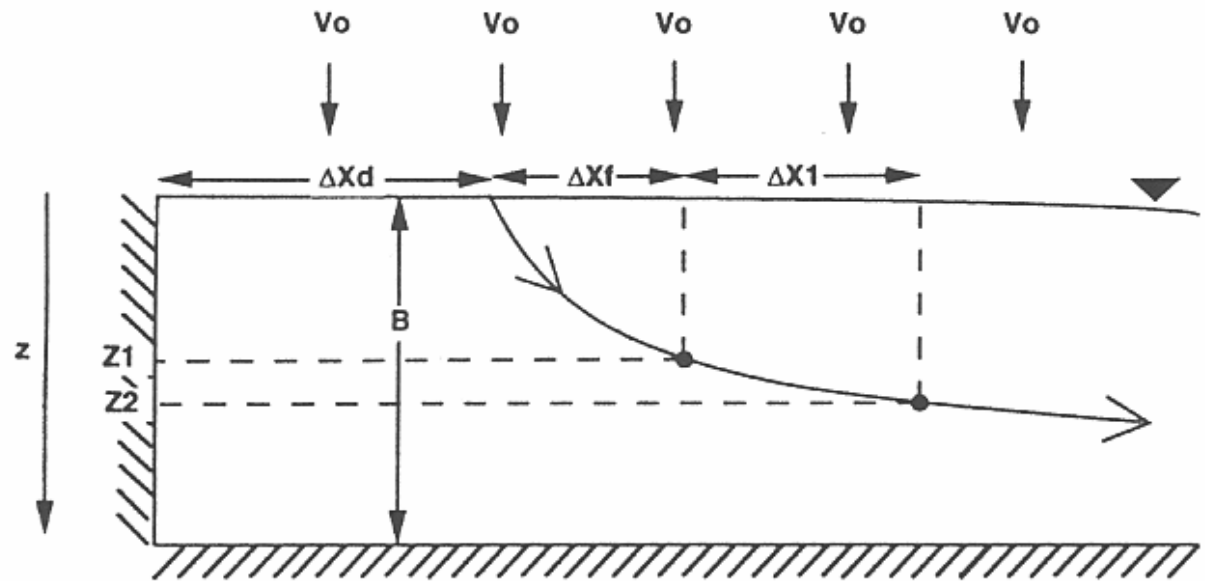
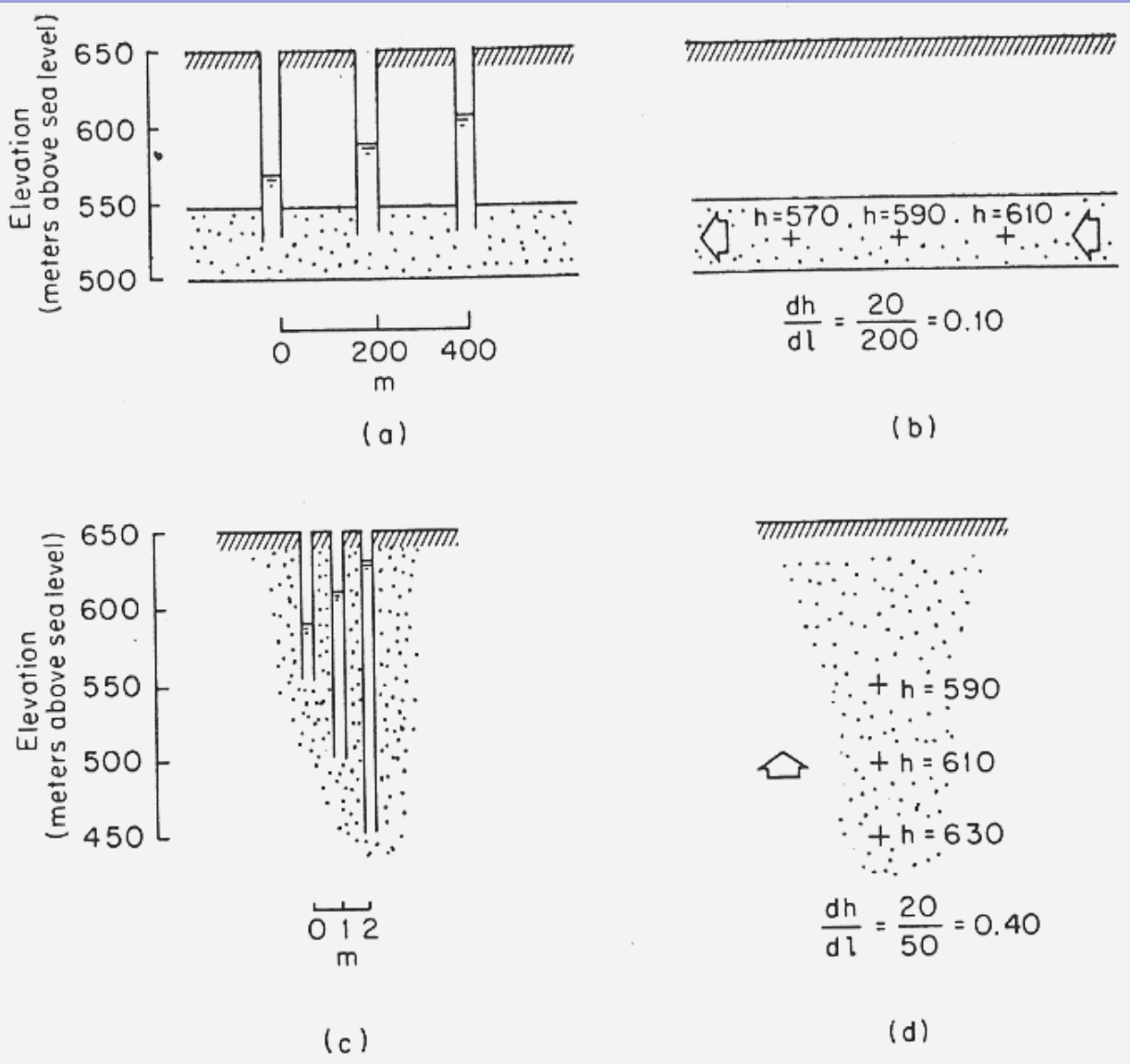


Fig. 6. Geometry of an unconfined aquifer used in the Vogel model.

Determination of hydraulic gradients from piezometer installations.



Steady flow through an unconfined aquifer resting on a horizontal impervious surface

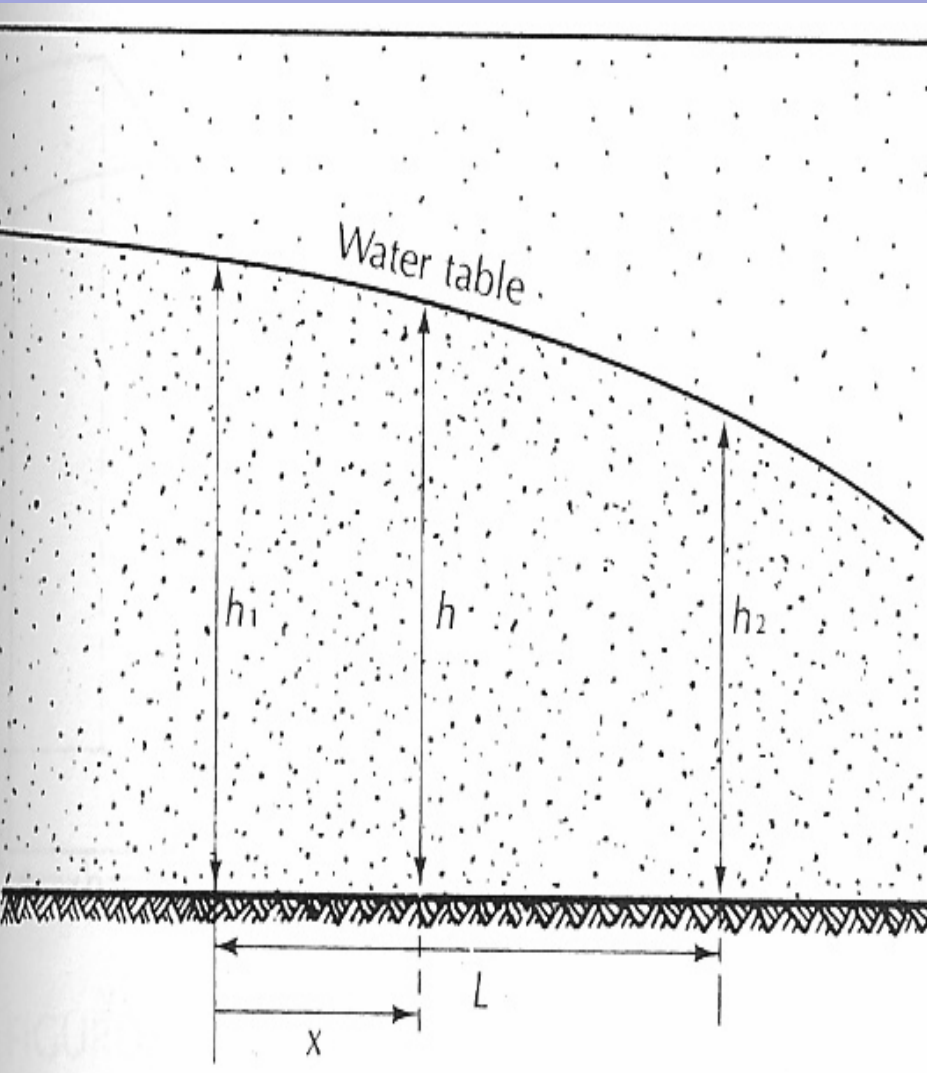


Figure 4.17. Fetter, *Applied Hydrogeology 4th Edition*

Steady flow through a confined aquifer of uniform thickness

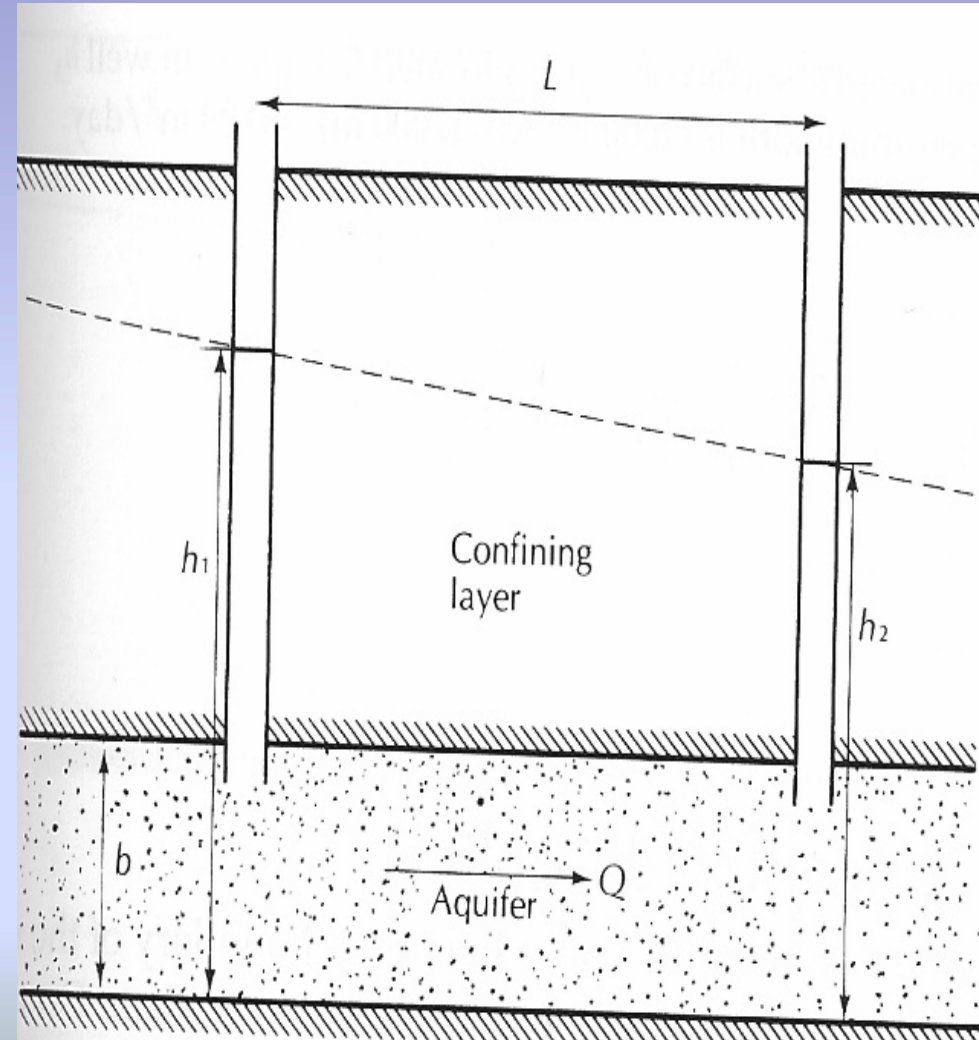
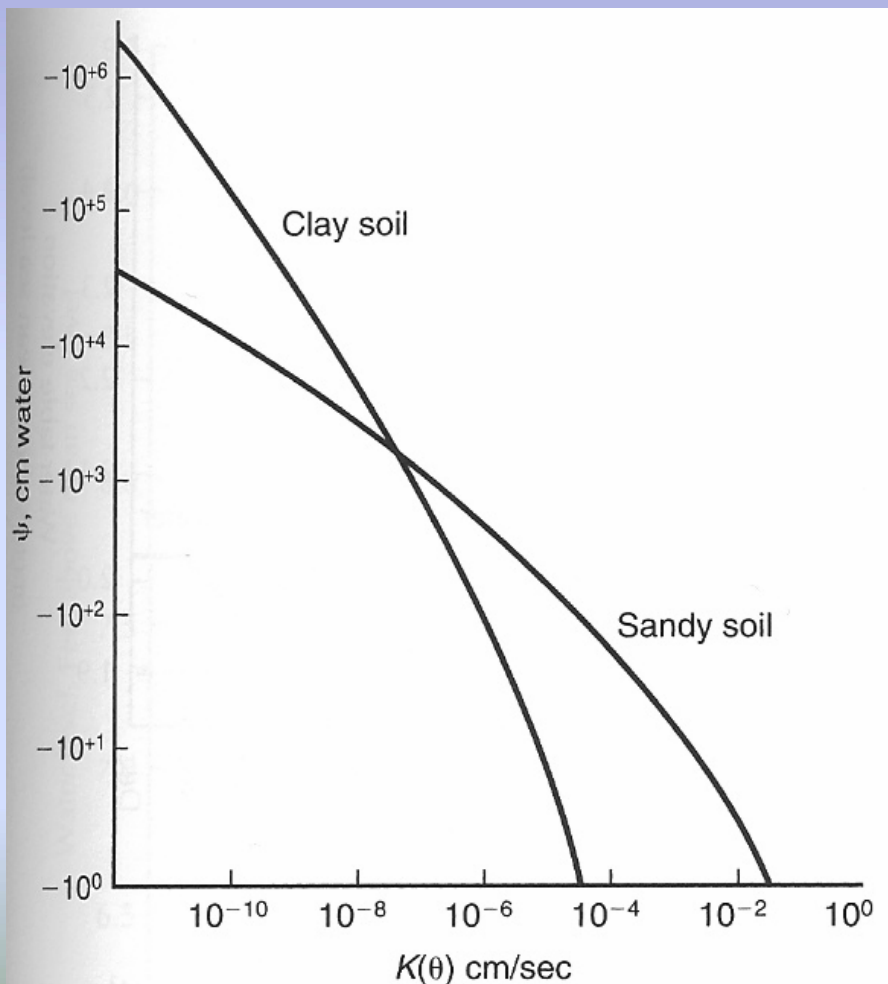


Figure 4.16. Fetter, *Applied Hydrogeology 4th Edition*

→ Equation that governs fluid flows are fairly straight forward. Only rigorous in interpretation and when in 3 dimensions.

→ Essentially, the velocity will depend on the hydraulic conductivity; therefore need to establish the hydraulic conductivity of the medium.



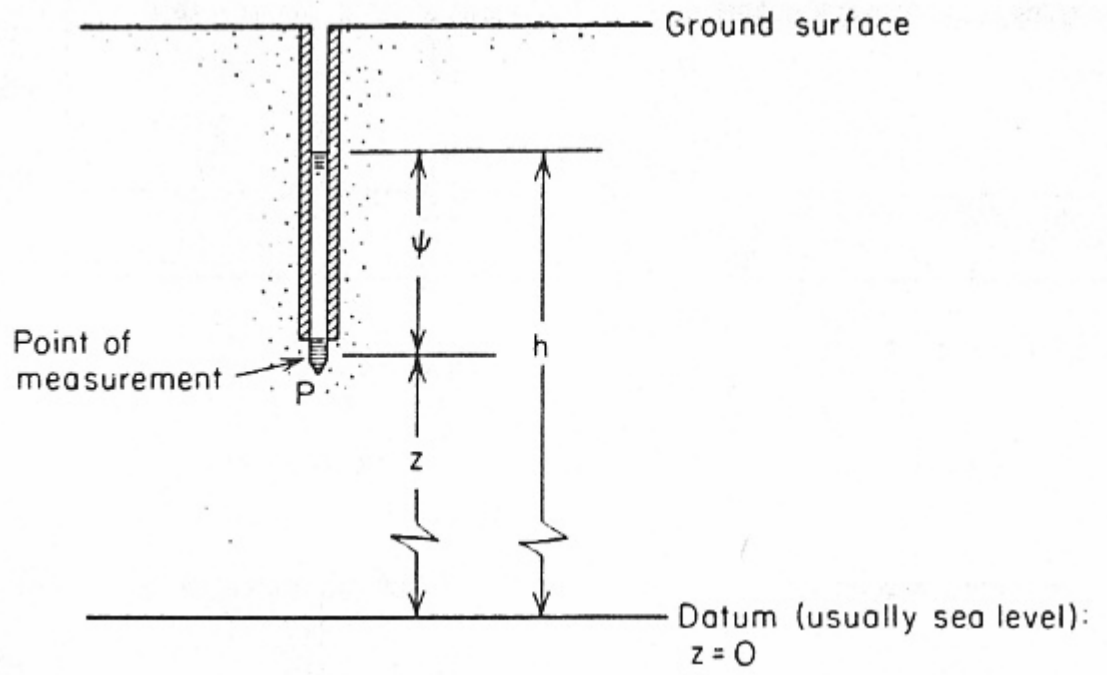
Typical soil-moisture-potential –hydraulic conductivity curves for a sandy soil showing the crossover effect for increasing moisture potential (decrease in water content.)

Figure 6.10. Fetter, *Applied Hydrogeology 4th Edition*

Measurement of hydraulic conductivity: (3 ways)

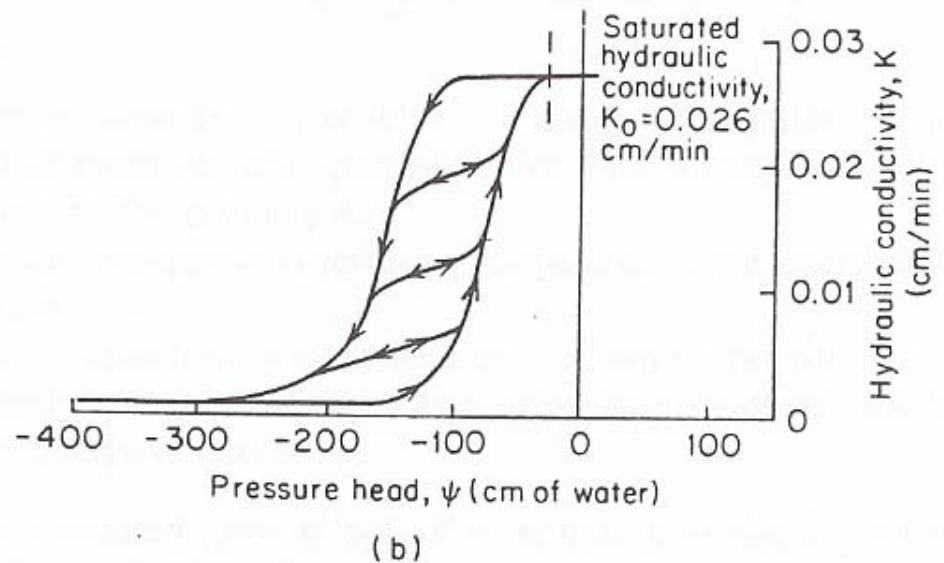
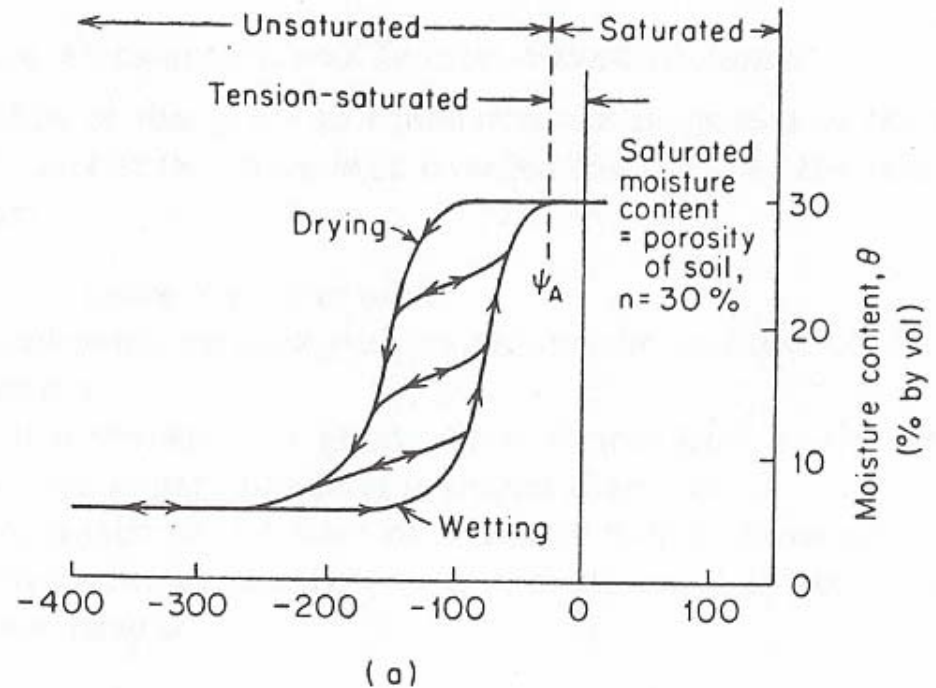
1. * Apply a hydraulic head and watch how the water flows through the sediments in the lab. One has a continuous supply of water and the other one drains in.

* Large range of potentials (log K). I.e. 2 orders of magnitude with peak at 10^{-5} . This is a log normal distribution-- Where graph peaks equals the average conductivity.



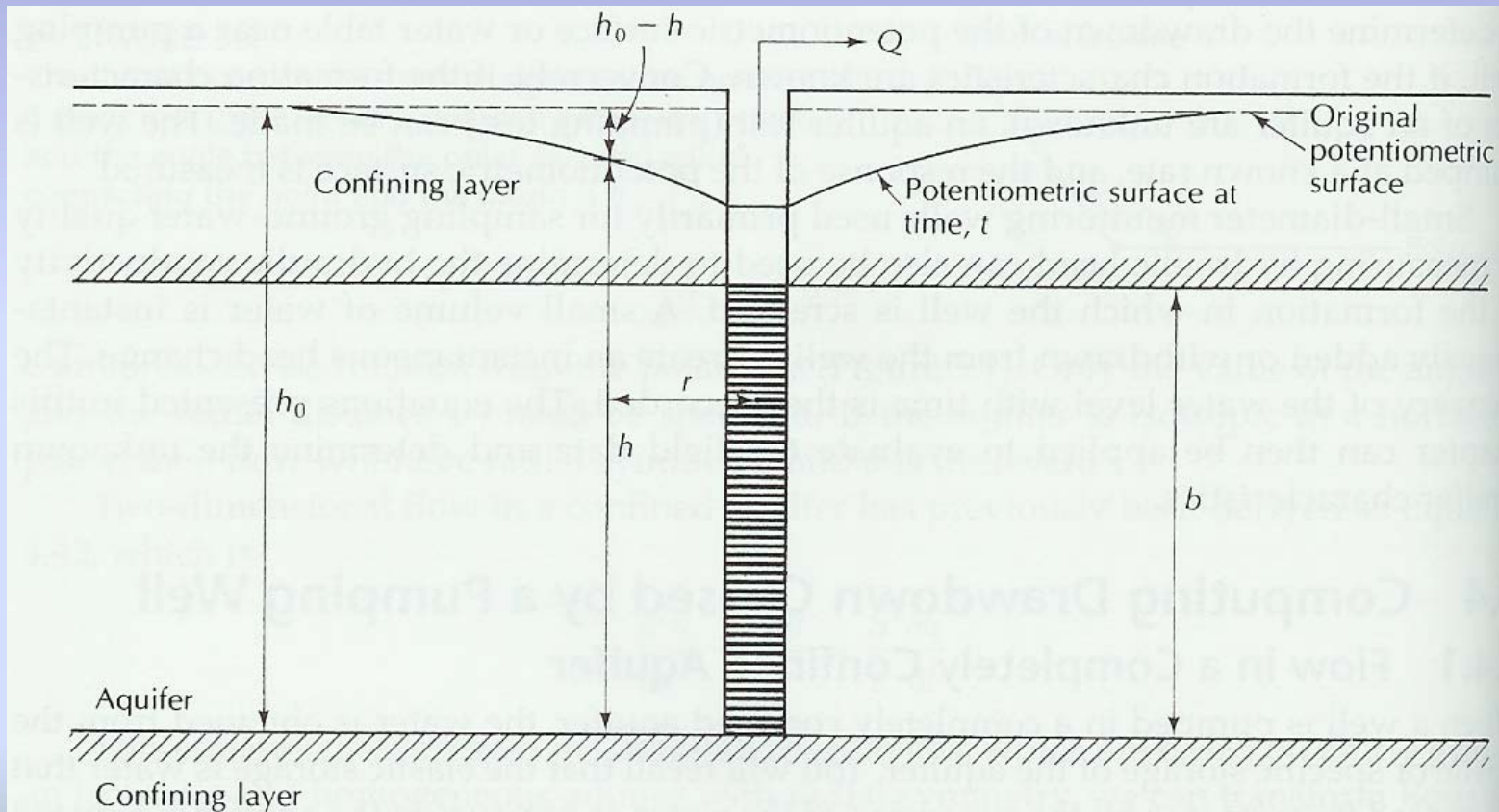
**Hydraulic head h ,
pressure head ψ , and
elevation head z for a
field piezometer.**

Characteristic curves relating hydraulic conductivity and moisture content to pressure head for a naturally occurring sand soil (after Liakopoulos, 1965a)



Measurement of hydraulic conductivity: (3 ways)

2) Pump testing: Water is pumped out and watched to find the rate at which it fills up. If the sediment is a clay, it will slowly return to equilibrium; if it is sand, it will return more quickly to equilibrium.



▲ FIGURE 5.2
Fully penetrating well pumping from a confined aquifer.

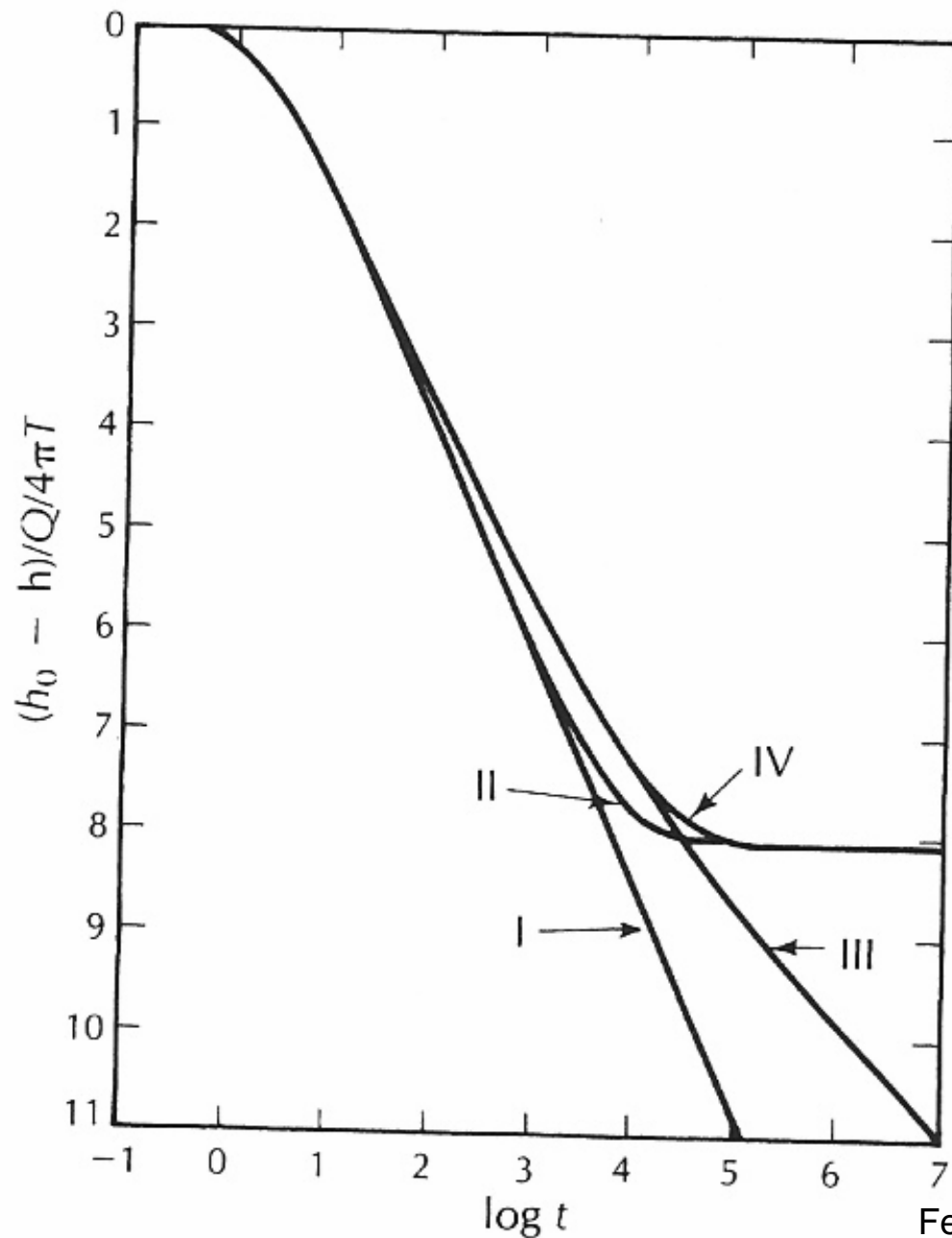
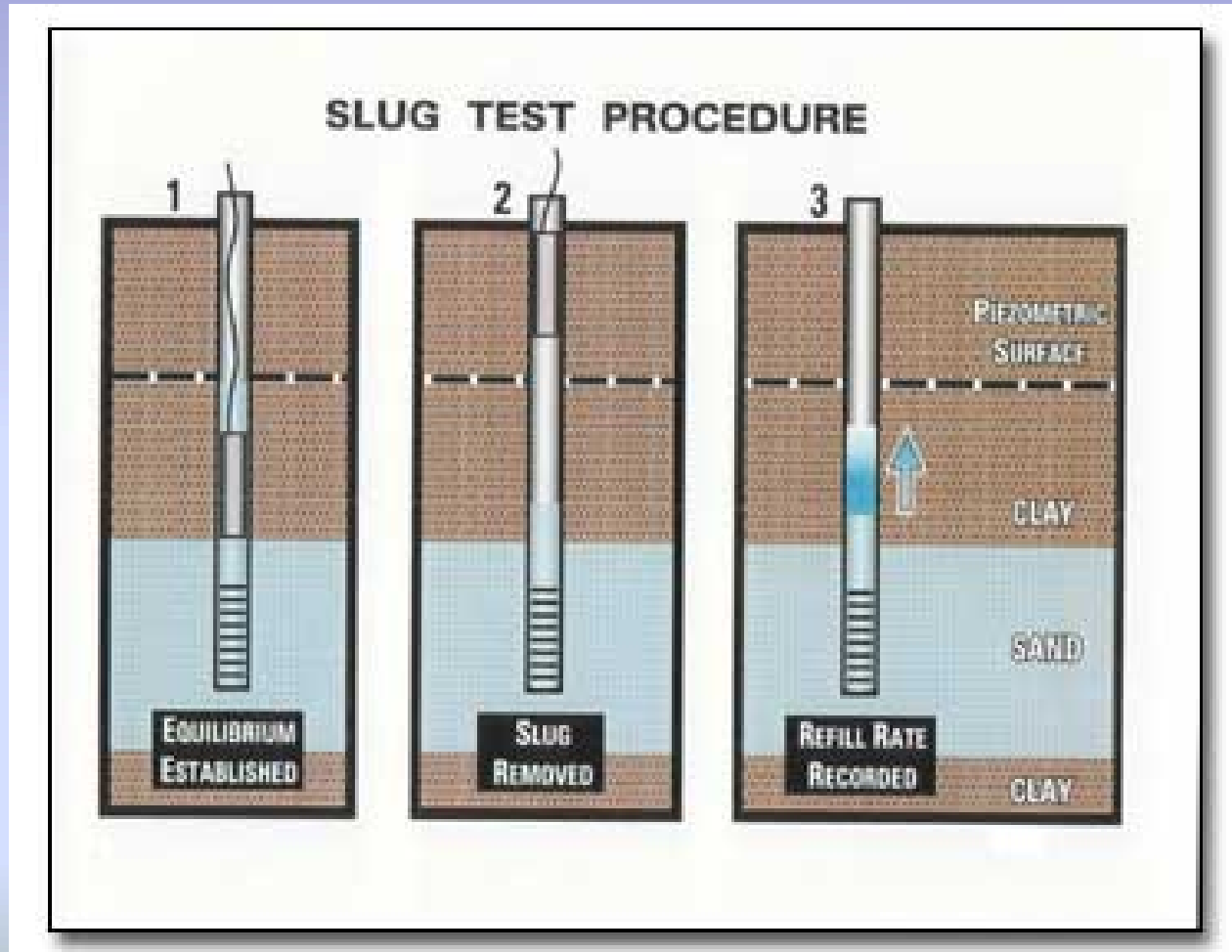


Figure 5.4 Plots of log of dimensionless drawdown as a function of log time for aquifer with various types of overlying confining layer.

- I No leakage
- II Leakage without storage in a finite semipervious layer
- III Leakage with storage in an infinite semipervious layer
- IV Leakage with storage in a finite semipervious layer

Measurement of hydraulic conductivity: (3 ways)

3. Slug test:



Hvorslov constructed an algorithm:

$$K = \frac{r^2 \ln(L/R)}{2L \Delta h}$$

Δh

r = well radius

L = length

• Can calculate standard K for a medium.

• Can also use a slug test where you drop a slug in to raise the surface of the water and watch the time it takes for the water to sink back down.

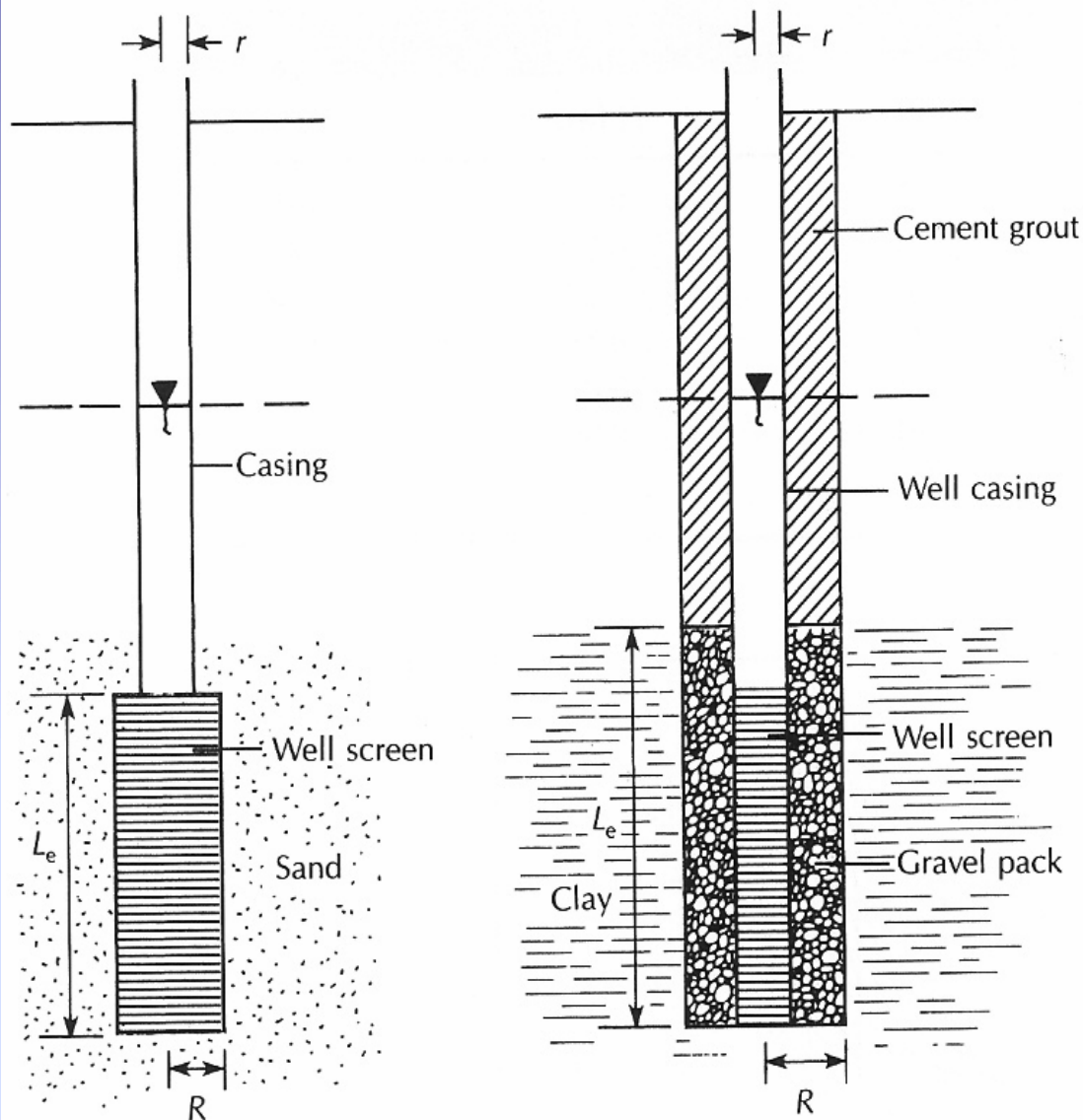
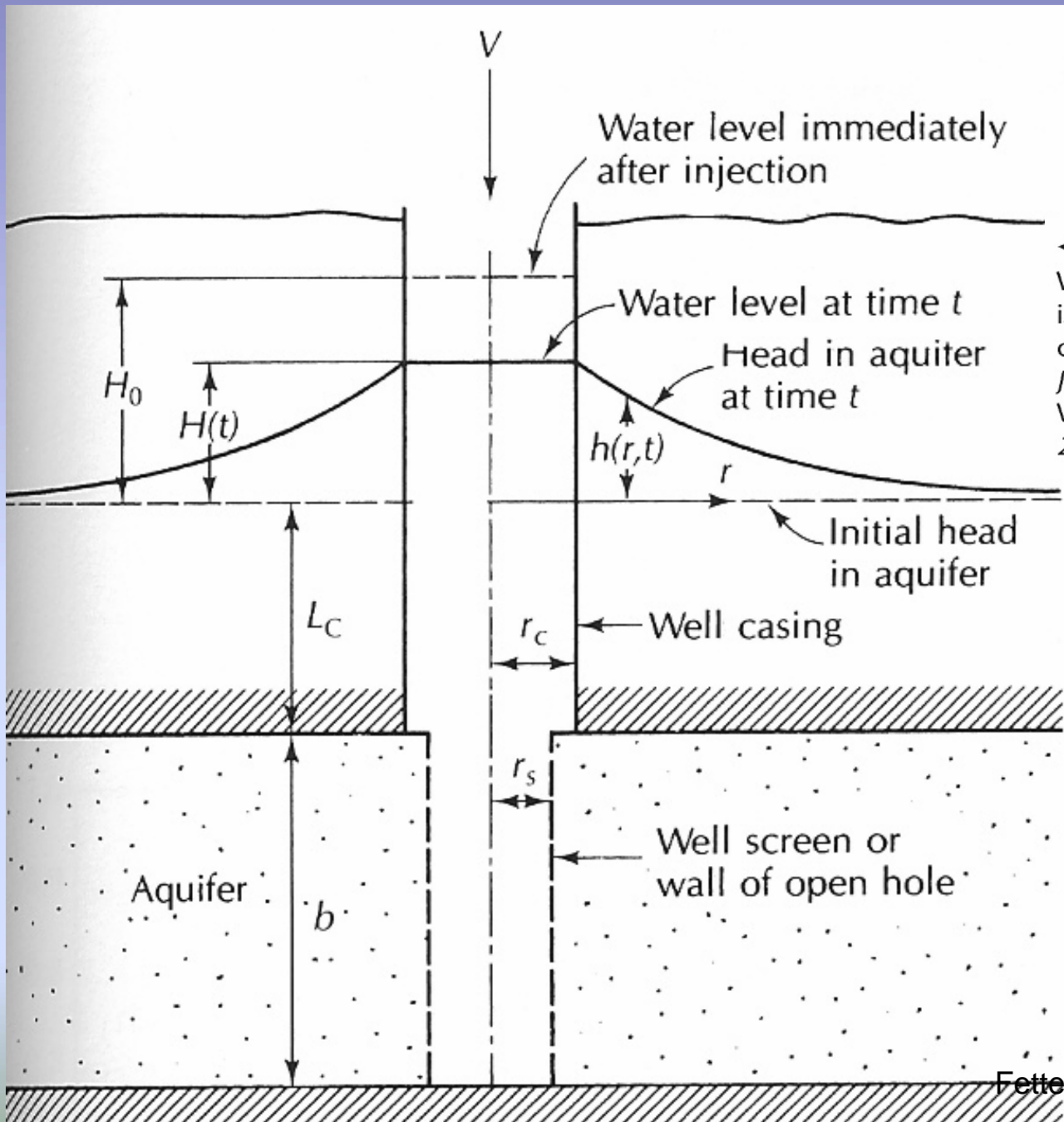


Figure 5.21. Piezometer geometry for Hvorslev method. Note that for a piezometer installed in a low-permeability unit the value R is the radius of the highest permeable zone that includes the gravel pack zone and L is the length of the gravel pack zone.



◀ FIGURE 5.18

Well into which a volume, V , of water is suddenly injected for a slug test of a confined aquifer. Source: H. H. Cooper, Jr., J. D. Bredehoeft, & I. S. Papadopoulos, *Water Resources Research* 3 (1967): 263–9. Used with permission.

→ All of these methods measure the horizontal conductivity → how fast it comes from the sides to recover.

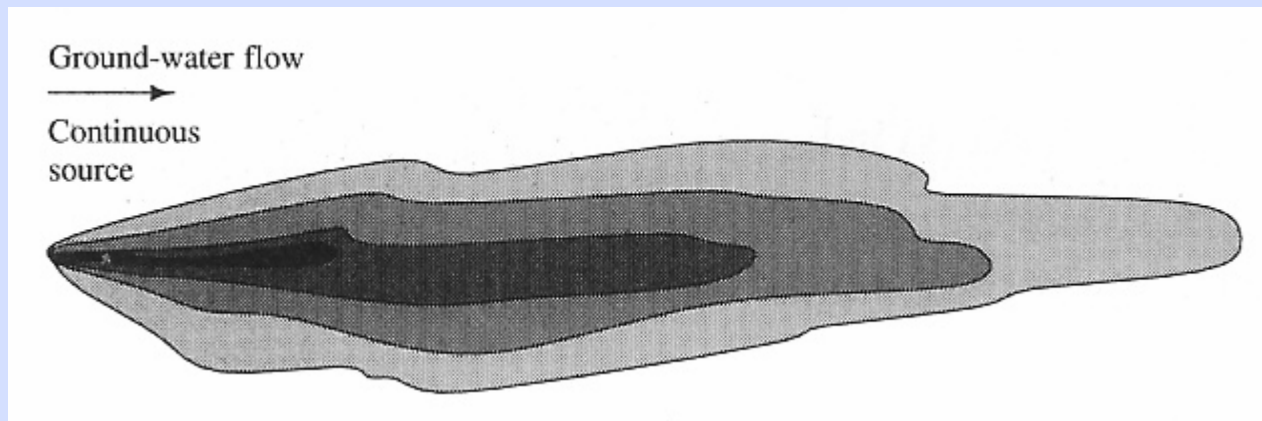
- **vertical conductivity is often quite different**; hard to get at by the way these pump tests are done.
- **Vertical is different because of anisotropy** → related to layering which can impede flow vertically but less so horizontally.

**$K_v = 1/10 (K_h)$ generally.
(h = horizontal).**

Other Conservative Groundwater tracers

- tracer moves at the same rate as the water
- not impeded by methods such as chemical reactions or degradation. Influenced only by physical processes such as mixing, diffusion, etc.

* See page from book: John Cherry, Camp Bordon, Ontario.



Plume resulting from the continuous injection of a tracer into a two-dimensional flow field.

Figure 2.11. Fetter, Contaminant Hydrogeology 3rd Edition

Injection of a slug tracer into a two-dimensional flow field shown at 3 time increments.

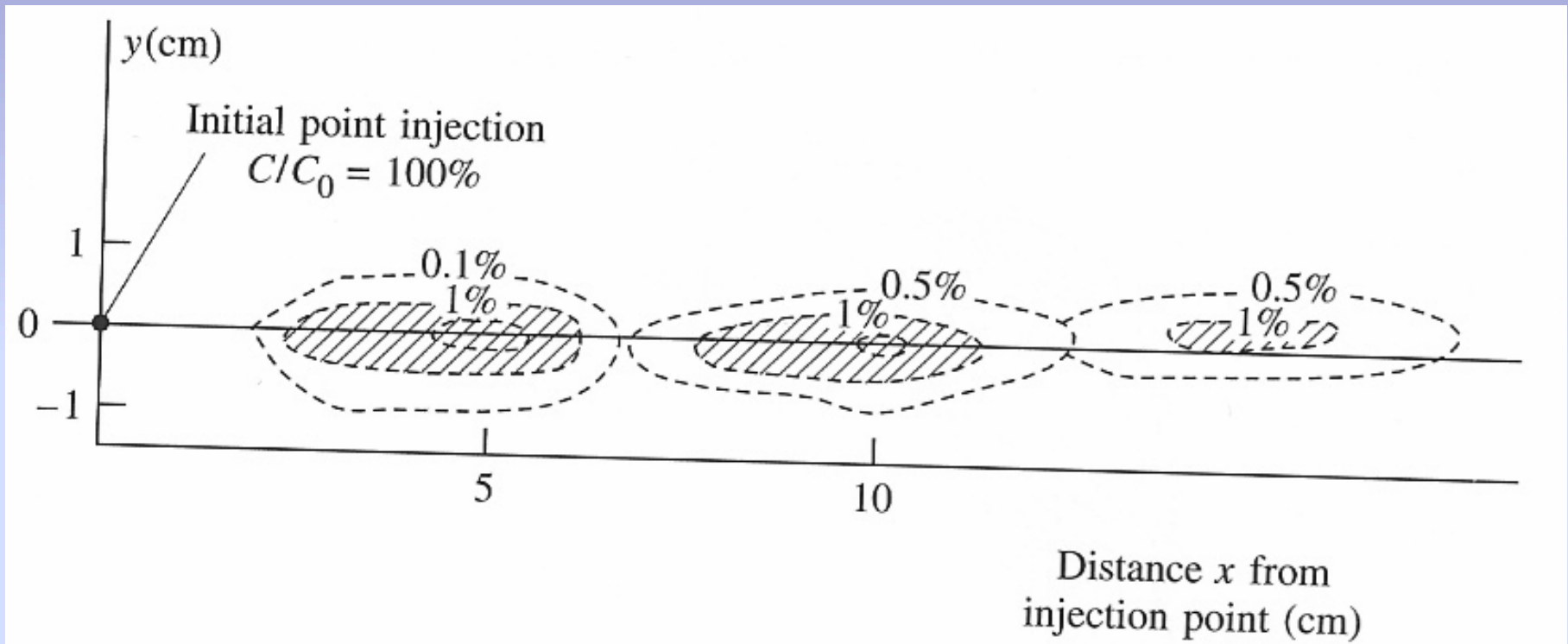


Figure 2.12. Fetter, Contaminant Hydrogeology 3rd Edition

- This was an old air force base turned into a hydraulic experimental station in Canada (this would never have been allowed in the US).
- The camp was established around 1940. A place was found to place the trash from the compound-- established a landfill in a relatively out-of-the-way place (without the strong safeguards of today).
- As the rain came down, it would leach out the components which would flow into the ground.

- Appears that the plume flowed from the landfill and moved from south to north (left of the flow in the picture is the ground water divide).

Distance/Time = Rate of Migration of the Plume.

- How things spread by diffusion.
- Important tracer and allowed understanding of system.
- Could compare gradient to hydraulic conductivities.

- Have been able to introduce other tracers into the systems-try to use benign tracers, organic solvents (build barriers to try to stop organic solvents).

→ This provided a good tracer:

1. Know when it started (1940's).
2. Produced a relatively benign set of contaminants,
mostly Cl⁻, which was simple and easy to analyze.

∴ Can trace K movement.

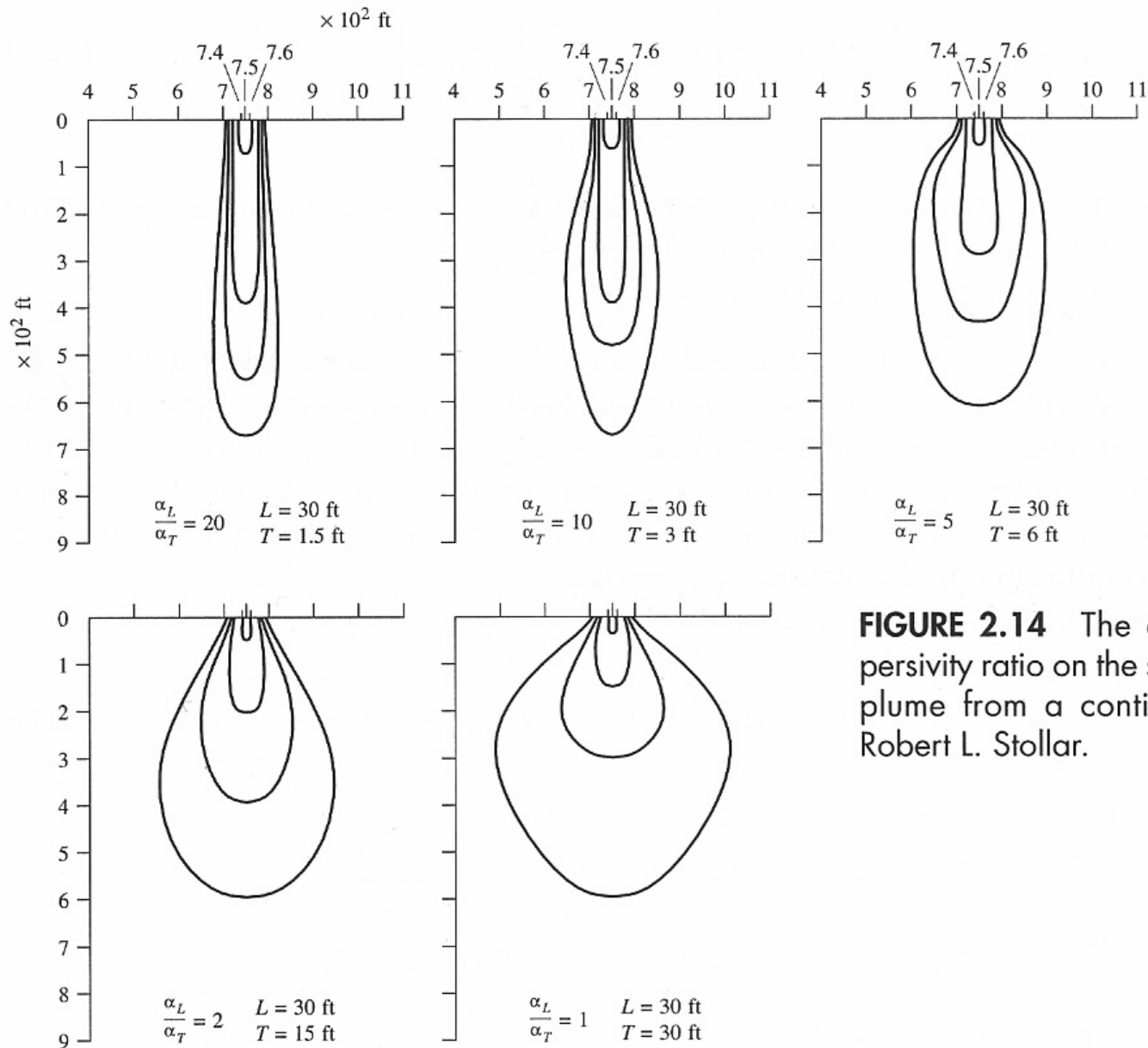


FIGURE 2.14 The effect of changing dispersivity ratio on the spread of a contaminant plume from a continuous source. *Source:* Robert L. Stollar.

Other tracers give time information

- ^{85}Kr → made from the reprocessing of nuclear fuel, power plants, high since 1950. But have to process about 200 liters of water, which is a substantial amount.
- CFC's → not conservative, consumed by organisms, adsorbed unto particles, but easy to measure.

^{85}Kr : radioactive form of krypton.

- formed during U and Pu fission.
- 52 dpm (0.9 Bq)/cc Kr.
- 1.58×10^{-11} $^{85}\text{Kr}/\text{Kr}$ in atmosphere.
- when at water table level, water amount of ^{85}Kr becomes equal to the atmospheric level- don't have to worry about solubility, only have to know the ratio $^{85}\text{Kr}/\text{Kr}$.
 - but need about 100L of water to count ratio, which can bias your results.
- man-made tracer

CFC's: CFC-11 → 220 ppt (today).
CFC-12 → 400 ppt (today).
– increasing at about 4%/year

Hundreds	Tens	Ones
#C - 1	#H +1	#F

Ex. CFC-11 → 1C, 0H, 1F = CCl₃F

CFC-12 → 1C, 0H, 2F = CCl₂F₂

- have a long residence time of greater than about 100 years.
- brings Cl into the atmosphere.

CFC-22 → 1C, 1H, 2F = CHClF₂

– short residence time- not as much Cl brought into the atmosphere-
so strive to produce this instead!

F-11 → 5.5 pmol/kg in atmosphere.

F-12 → 2.3 pmol/kg in atmosphere.

Problems:

- temperature dependence : cold- dissolve more gas; warm - dissolve less gas.
- anthropogenic sources (landfills).
- organic solvents dissolve large amounts of F-xxx in the air ---> to ground water.
- non-conservative in its behavior because it is organic- can be eaten or absorbed (only F-11 is eaten).
- F-113 - behaves similarly to “nasty” molecules but has F instead of H.

Travel time? "Conundrum of Solomon"

- In reality are lines that are almost totally horizontal (paralleled by tritium-helium age).

** Very important for plume hydrology.

(Keep in mind that no flow is coming out of the bottom of the box).

The vertical velocity is not constant, higher at the beginning and lower further away, a function of z .

[This is happening at Cape Cod, didn't realize how fast the water was moving until it got into the water supply wells].

Behaviors of contaminants

I. Movement:

- the release (plume) will not spread at a constant rate because there are different paths that it could take- it moves with the water- will take short or long paths around grains.

II. Two ways organics migrate

- 1) **Dissolved:** solubility is proportional to mobility (high solubility, then high mobility).
- 2) **NAPL:** non-aqueous phase liquid

NAPL: non-aqueous phase liquid

A) * **L-NAPL's**: less dense than water.

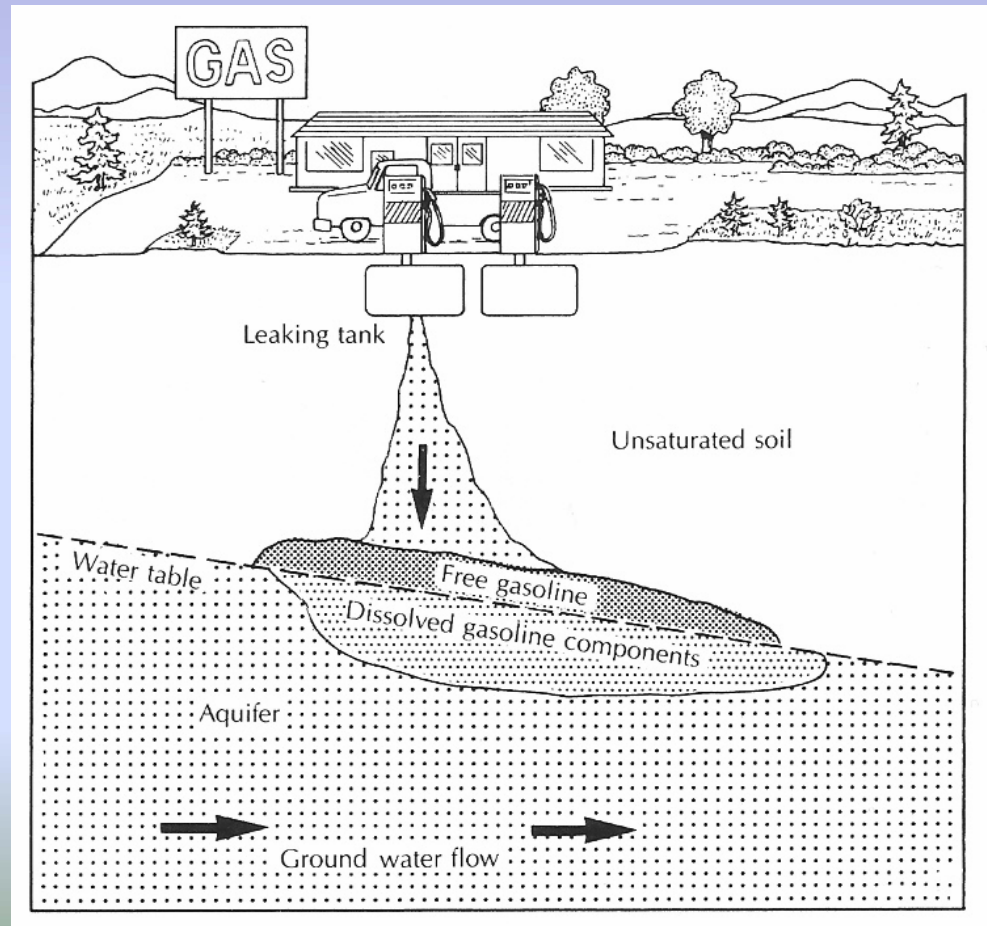
–ex. *Gasoline* - forms a pocket which floats on the water table

–soluble material such as *benzene, toluene, xylene* (all organics) dissolves in the water

→ the gas just evaporates.

Organic liquids such as gasoline, which are only slightly soluble in water and are less dense than water, tend to float on the water table when a spill occurs.

occurs. Figure 10.19. Fetter, *Applied Hydrogeology 4th Edition*



NAPL: non-aqueous phase liquid

B) * **D-NAPL's**: denser than water.

–ex. *methylene chloride* - when its spilled, it is gone and not seen again.

–*trichloroethylene* - breaks down to worse things.

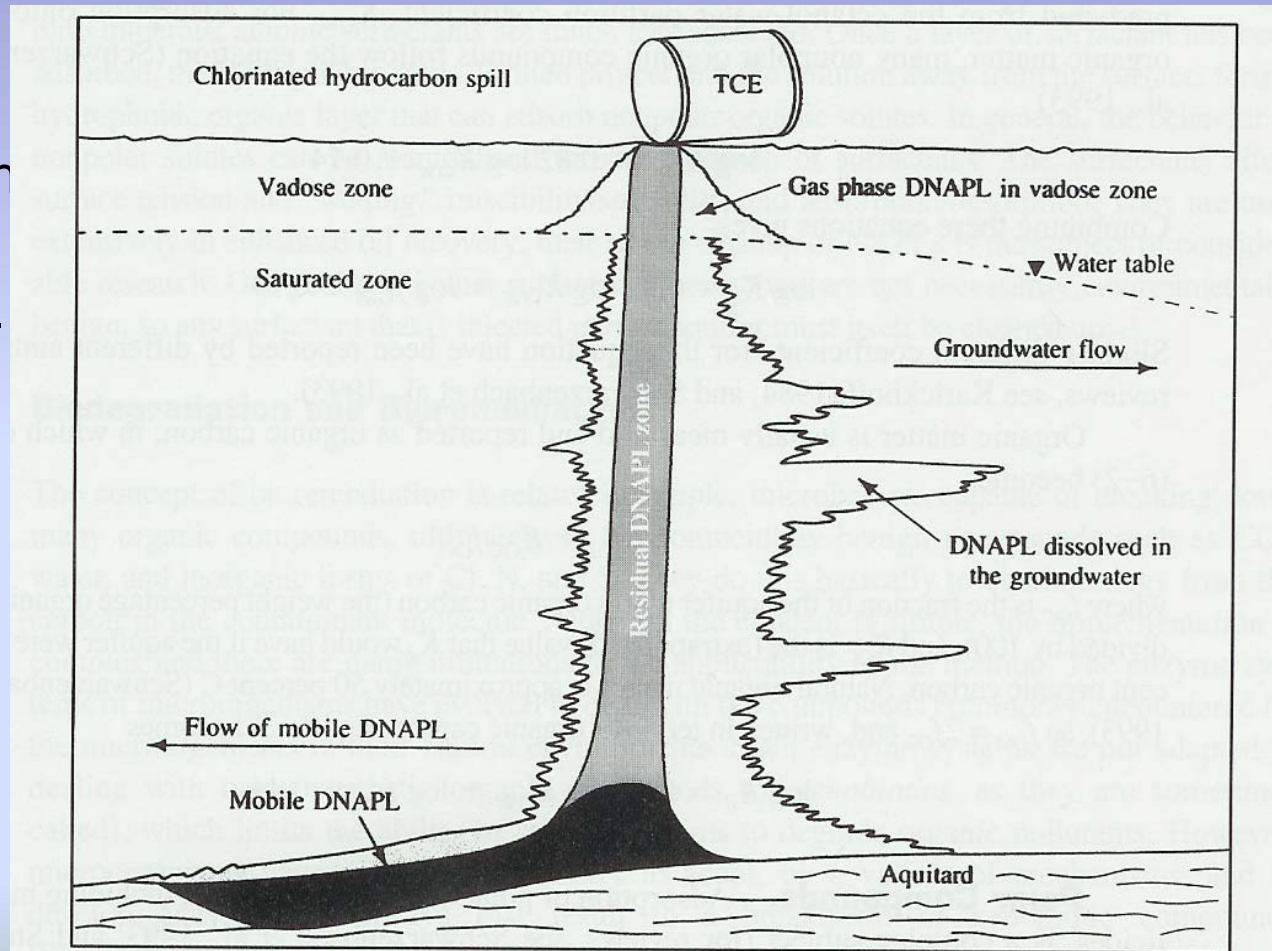


FIGURE 6-7 General distribution of a DNAPL in the subsurface following a spill. CONTAMINANT HYDROGEOLOGY by Fetter, C. W., ©1993. Reprinted by permission of Prentice-Hall, Inc., Upper Saddle River, NJ.

III. Two types of contaminants

(Very broad classes)

A. DNAPL (Dense Non-Aqueous Phase Liquid)

- * TCE (density = 1.46) Solubility in water: 1180ppm
- * TCA (density = 1.33) 1180ppm
- * PERC (perchloroethylene, density = 1.6)
- * Methylene Chloride (density = 1.33) 1300ppm
(highly soluble)

→ These are extremely dense.

B. LNAPL (Light Non-Aqueous Phase Liquid)

- * Benzene 1800ppm
- * Toluene 500ppm
- * Vinylchloride (density less than water and highly volatile)

***NOTE: solubilities given by weight (1 ppm = mg/L).

➔ The drinking water standard for these contaminants is at 5 ppb (parts per BILLION)

C. Characteristics:

1. Densities of NAPL's range from .8x water to 1.6x water
2. Density is a function of the chloride.
3. Higher density contaminants will move quickly down a water table aquifer.
4. Porous vs Fractured Systems exhibit very different behavior.
5. Pure phase vs. dissolved phase → different problems associated with each.

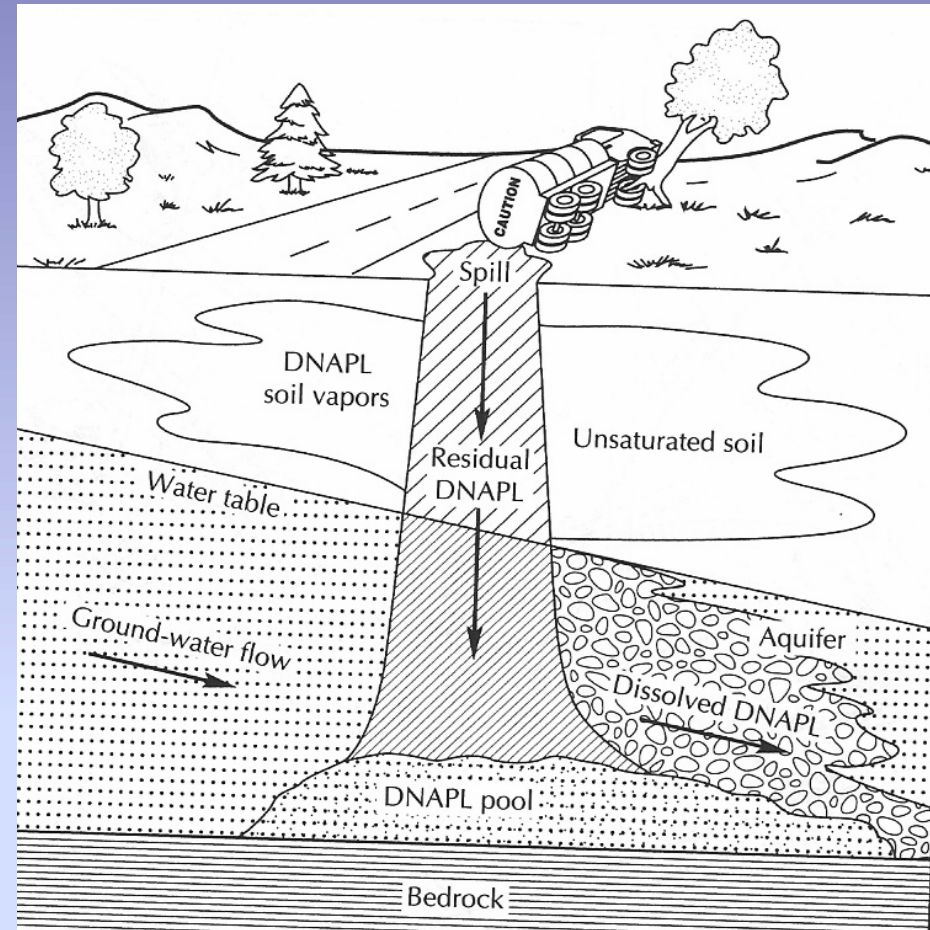
IV. Example: Spill of trichloroethylene (density = 1.33 g/cc).

- not that soluble.

→ What happens: migrates down to bedrock and pools there & slowly dissolves as water flows past.

-- OR --

- if there are lenses of clay, it will pool on them.
- distributed over a wide area, it is impossible to find it and get it out!
- try to pump it out → only end up dissolving it because the water flow over organics increases.
- if spill is 100,000 L at a concentration of 5 ppm, then you have to pump 2×10^9 L of water.
- concentration decrease due to less water-- when the pumping stops, the water flow returns to normal and the concentration goes up.



Organic liquids such as trichloroethylene, which are only slightly soluble in water and are more dense than water, may sink to the bottom of an aquifer when a spill occurs.

Figure 10.20. Fetter, *Applied Hydrogeology* 4th Edition

V. Remediating systems

→ Is there really the possibility of remediating organically contaminated systems or are funds better spent elsewhere?

A. Questions:

1. size of the spill in terms of 1000's of liters of pure contaminant - need to remediate to a few ppb or less??
2. mobility. How fast will this move?
3. degradation → rate at which organic material will degrade is dependent on the medium through which it travels. i.e. Small amounts of organic carbon is enough to slow this down.

Remediating systems

B. Problems:

Volume calculations:

- 10,000 liters of contaminant dissolves to affect 6,000,000,000 liters of plume!!
- 15,000 liters of contaminant dissolves to affect 40,000,000,000 liters of plume!

(This takes less than a tanker to contaminate this much!)

How much pure product is there and how can we get to it?

–This ends up causing a lot of problems especially since water wells are close.

How hard will it be to remove ie 5 billion liters of water? **EXTREMELY hard!**



Remediating systems

C. Remediation by the Pump and Treat Scenario

- Process relies on removing large volumes of water. This is not an equilibrium process. Natural flow may be 50 m/year and the flow with the pump may be 500 m/year.
- Find that after the water seems clean and the pumps are stopped, that the concentration of the contaminant will rise again.
- In a past situation with extensive contamination, 1 million liters were drawn out of the well and treated.

1mg/L → trivial 1 kg.

The consultant even got out 30 kg. But the spill contaminated on the order of 30 million kgs!!

** This method seems to be the accepted way to clean spills, although it is not necessarily efficient.



Remediating systems

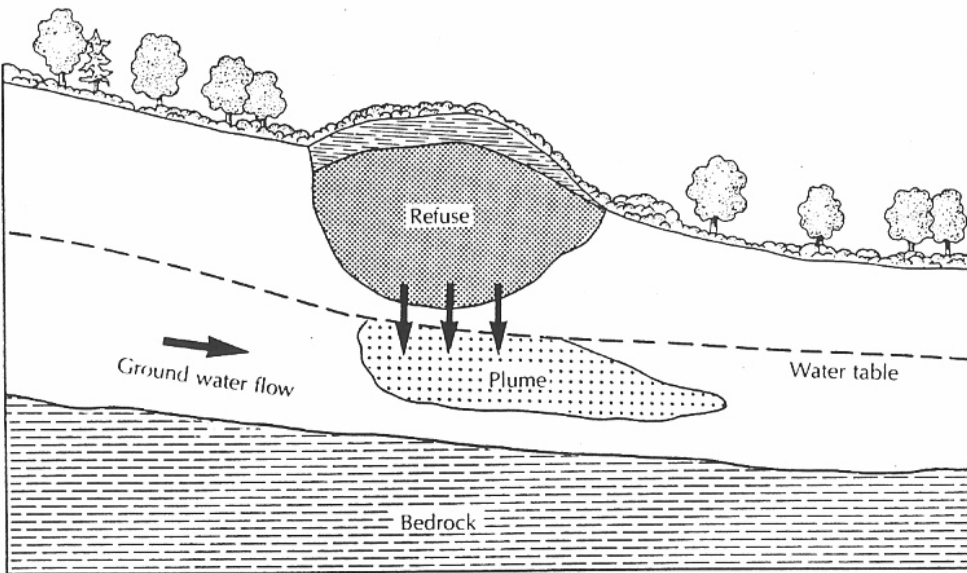
D. Why won't it work?

- * No aquifer is perfectly homogenous -- Even the simplest aquifer is heterogeneous (ie grain size).
- * Conductivity varies by at least an order of magnitude ie about 1 log unit.
 - What does this do to the system?
 - → Small lenses form whose conductivity is 1 or 2 orders of magnitude less than the rest of the system. The NAPL will congregate on the low permeability area. Therefore it doesn't fall neatly straight down, but will cascade down and form a complex distribution.

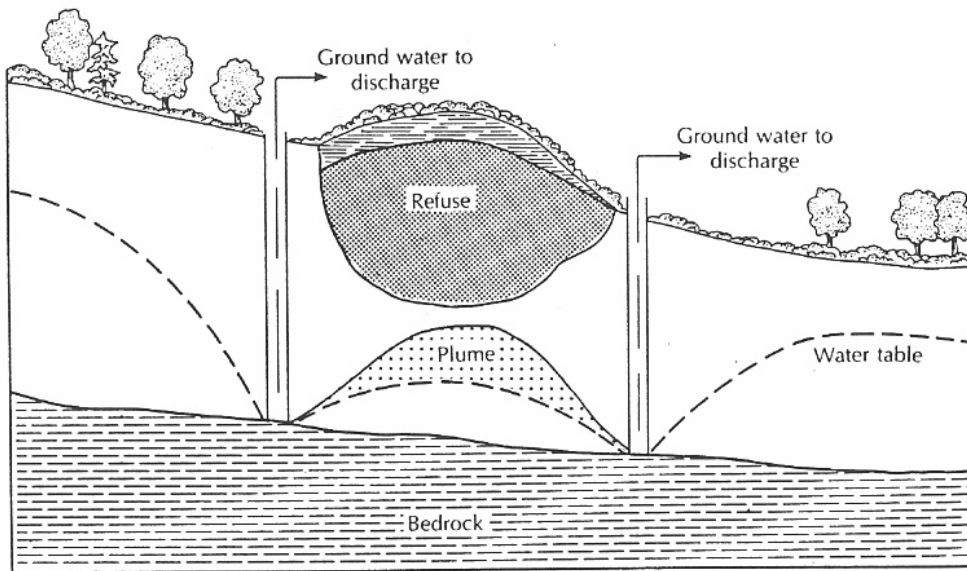
Remediating systems

D. Why won't it work?

- Water will not be seeing NAPL and therefore may take a long or infinite amount of time to pump out. More time is needed for higher heterogeneity → These stringers of low conductivities hold NAPL and don't allow solubility and movement of it → never attack the material in fine-grained, low conductivity sediments.
- * L NAPL is easier to extract by vapor extraction if not too much is dissolved in the plume.



Before pumping



After pumping

Use of Extraction Wells to Remove Contaminated Ground Waters

FIGURE 10.27 Use of extraction wells to remove contaminated ground water. Source: U.S. Environmental Protection Agency.